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Technosols designed for rehabilitation of mining activities using mine spoils and biosolids. Ion mobility and correlations

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30 **Abstract**

31 The restoration technologies in areas degraded by extractive activities are more efficient
32 under the use of their own spoils. Reducing deficiencies in physical properties, organic
33 matter, and nutrients with a contribution of treated sewage sludge is proposed. This
34 experiment was based on a controlled study using columns. The work was done with
35 two limestone quarry spoils, both very rich in calcite. Two biosolids doses were
36 undertaken (30,000 and 90,000 kg/ha of sewage sludge) in addition to a different quarry
37 spoils used as substrates. The water contribution was provided by a device simulating
38 short duration rain. The leached water was collected 24 hours after the last application.
39 Nitrate, ammonium, phosphate, sulfate, and chloride ions were determined, as well as
40 the pH and electrical conductivity. The electrical conductivity limit value is <1000
41 $\mu\text{S}/\text{cm}$. These values will be met from the fourth irrigation application onward, while
42 the values up to that point were far superior. Significant nitrate concentrations appeared
43 that may pose an environmental contamination risk. A comparison between the
44 concentrations of the chemical elements obtained in the leachates from our experiment
45 and the established limit values for water of the third quality group has been performed.
46 The electrical conductivity correlated well with the cations, with the exception of
47 potassium. For sulfates, significant correlations were obtained with the Mg^{2+} , Ca^{2+} , and
48 K^{+} cations. The chlorides showed excellent correlation with the sodium.

49

50 **Keywords** Ion mobility • Irrigation • Quarry spoils • Sewage sludge

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56 **1 Introduction**

57 The restoration of extensive areas degraded by mining activities requires the use or their
58 own waste materials (Jordán et al. 1998; Tedesco et al. 1999; Ram et al. 2006; Jordán et
59 al. 2006; Almendro-Candel et al. 2014, Galán et al. 2014). These materials do not
60 possess the necessary fertility to ensure a successful process of restoration
61 (implementation of adequate plant cover). Therefore, it requires the addition of organic
62 amendments to achieve efficient substrate (Jordan et al. 2008). The obligation to restore
63 abandoned quarries and the correct application of biosolids is guaranteed by the
64 legislation on waste management, biosolids and soil conservation (Jordán et al. 2008).
65 Technosols are one of the latest additions to the World Reference Base for Soil
66 Resources (FAO 2006). This new reference soil group contains a large range of artifacts
67 and materials of natural and anthropic origin. They include a variety of refuse-based
68 soil-like quarry spoils, landfills, ashes, or sludges, whose properties and pedogenesis are
69 dominated by their technical origin (FAO 2006). An adequate Technosol selection,
70 based on its nature and intrinsic properties, can constitute a valuable and cost-effective
71 solution for soil remediation and waste management (Novo et al 2013). Sewage sludge
72 application in restoration has demonstrated its efficiency in previous studies (Clapp et
73 al. 1986; Albiach et al. 2001; Pond et al. 2005; Jordán et al. 2008; Soriano-Disla et al.
74 2014). The use of treated sewage sludge can be a guarantee of success in the restoration
75 of areas affected by mining activities, but it is important to preserve the conservation of
76 the environment with less risk of contamination of surface and groundwater. Many
77 physical and chemical properties in soils amended with sludge, such as water retention
78 capacity, aggregate stability, contribution of N, P, and other nutrients to crop growth,
79 depend, to some extent, upon the quantity of organic matter in the sludge that is added
80 (Roldán et al. 1994). Knowledge about the quantity of organic matter in the sludge can

81 be used to estimate the quantities that must be applied to the soil (Giovannini et al.
82 1985). Use sewage sludge and quarry spoils to construct Technosols represent an
83 innovative strategy of waste management, whose application allowed the species to
84 grow and develop (Novo et al. 2013). The materials resulting from the acquisition of
85 arid particles produced from crushed limestone present very limited results in
86 restoration because of their chemical and mineralogical characteristics. The materials
87 resulting from limestone extraction, with elevated soil stoniness and low fertility, are
88 usually used in restoration, and the efficiency depends upon the organic corrections
89 (Jordán et al., 2008).

90 The studied area, Sierra de Callosa has an area of 8 km² and a maximum altitude of 566
91 masl (Jordán et al. 2008). Geologically, it is found to be primarily comprised of
92 carbonated rocks. Limestones have been worked for many thousands of years, initially
93 for building stone and agricultural lime and more recently for a wide range of
94 construction and industrial uses. In former days, limestone quarries were small
95 operations with limited production due to the labor-intensive nature of the business and
96 restricted markets for stone. Stone quarries can answer the demand of aggregates for
97 concrete applications (Jordán et al. 2008).

98 The main objective is to evaluate the ion mobility and correlations in columns packed
99 with quarry spoils from a limestone quarry amended with biosolids.

100

101 **2 Materials and methods**

102 **2.1 Substratum**

103 The experiment was carried out using two different limestone quarry spoils, both very
104 rich in calcium carbonate (450-750 g/kg). The first, of poor quality, originates from the
105 crushed limestone (Z). It is composed of coarse materials (up to 75 % by weight) and

106 sand. The other tested waste material comes from the extraction of limestone. This
107 waste was formed by the levels of interspersed non-limestone materials and remains of
108 stripped soils (D). This usually presents more balanced textures but with elevated
109 heterometry soil stoniness (up to 60 % by weight), and is richer in clays (approx. 25 %
110 by weight). Analytical parameters were determined in accordance with ISRIC (1993).
111 Organic matter was quantified by wet oxidation, using the method of Walkley and
112 Black (1974); total nitrogen by the Kjeldahl method; pH in a 1:2.5 soil/water
113 suspension; the CaCO₃ equivalent with a Bernard calcimeter; by atomic absorption (Ca,
114 Mg, Fe, Cu, Mn, Zn) and flame atomic emission spectroscopy (Na,K). Particle size
115 distribution was determined by the pipette method.

116 The characteristics of the mineral substrata employed appear in Table 1. Both these
117 materials were amended with the biosolid according to Alcañiz et al. (1997) quarry
118 restoration methodology.

119

120 **2.2 Sewage sludge**

121 The biosolid used in this experiment comes from a wastewater treatment plant located
122 near Aspe (Alicante). Prior to the composting process, the sludge needs to be mixed
123 with a bulking agent, a supporting structure that favors aeration, absorbs humidity, and
124 furthermore contributes with organic matter. Chopped hay and sawdust are used as
125 bulking agent, and silos exist for their storage. Hay favors aeration, sawdust absorbs
126 humidity, and both materials constitute sources for carbon. The composition by volume
127 of the sludge-bulking agent mixture is 50 % sludge, while the remaining 50 % is 1/4 hay
128 and 3/4 sawdust. This sludge-bulking agent mixture progresses through the composting
129 tunnel and is simultaneously homogenized by a tumbler, which in addition to permitting
130 the progress and homogenization of the mixture, promotes its aeration. During the first

131 weeks, the mixture is placed upon a porous base connected to an air injection system
132 using fans or blowers, which maintains discontinuous forced aeration. Afterwards, the
133 aeration is passive and natural (Clapp et al., 1986; Hernández-Fernández, 1986).
134 For the biosolid analysis, total content of metals was determined following microwave
135 digestion using HNO₃ and analyzed by inductively coupled plasma mass spectrometry.
136 In the solution thus obtained, the solubilized elements except for nitrogen were
137 assessed. This was determined by the Kjeldahl method, which quantifies the organic
138 nitrogen and ammonium contents within the sample. The easily oxidizable organic
139 carbon was estimated by sulfochromic digestion and subsequent assessment with
140 Mohr's salt, while the easily oxidizable organic matter was calculated by multiplying
141 the organic carbon by 1.72. The total organic matter was obtained by calcination in a
142 muffle furnace at 500 °C for 2-4 h. Table 3 shows sewage sludge composition.

143 **2.3 Columns**

144 The experiment was based on a controlled study using columns. Fifteen columns, each
145 30 cm tall (Fig. 1), were constructed from 10.5 cm internal diameter PVC pipe that was
146 cut into two 15 cm lengths. Each column was divided into two different 15 cm sections,
147 the first one from 0 to 15 cm and the second from 15 to 30 cm. For each treatment three
148 replicates were done (Table 1).

149

150 **2.4 Treatments**

151 Two treatments and a control were applied (30,000 kg/ha and 90,000 kg/ha), which
152 depended upon the quantity of sludge applied and the experimental design (Table 3).
153 The sludge was applied on the surface and mixed with the soil, simulating a plowing or
154 tilling action, producing a homogenous mixture within the uppermost 15 cm of soil
155 (Almendro et al., 2007; Almendro et al., 2014).

156

157 **2.5 Irrigation**

158 In order to establish the closest parallels between real conditions and those of the
159 experiment, the soil contained in the columns was irrigated (eight applications) using
160 tap water. The first five irrigations occurred every two weeks and the last three once per
161 month. The irrigation applications lasted six months. Collection of the leached water
162 was carried out 24 hours after the last application. This irrigation is equivalent to
163 weekly rainfall of 100 mm (rainfall conditions during times of maximum abundance).
164 The contribution of water was provided by a device that simulated short rainfall or a
165 flood irrigation system that covered the surface and then percolated into the soil
166 (Almendro et al., 2014). It consists of a plastic recipient with holes punched in the
167 bottom (Fig. 1).

168 In order to control what was incorporated into the soil columns by the irrigation, water
169 samples were taken from each column. Irrigation water characteristics were determined
170 first (Table 4), i.e., pH, electrical conductivity, the Na^+ , K^+ , Ca^{2+} , and Mg^{2+} cations, as
171 well as the Cl^- , SO_4^{2-} , NO_3^- and PO_4^{3-} anions (Cánovas 1980). The parameters analyzed
172 in the leachates followed the Standard Methods (APHA Standard Methods, 2005). Na,
173 K, Ca, Mg were determined by atomic absorption spectrometry following acid digestion
174 in a microwave. Cl^- by titration by the Mohr method; SO_4^{2-} based on the formation of a
175 colloidal form with barium; soluble PO_4^{3-} using the vanadomolybdophosphoric acid
176 method.

177

178 **2.6 Leachates**

179 The electrical conductivity was determined by an electrical conductivity meter, which
180 incorporates a conductivity cell, considering 25 °C as the reference temperature,

181 according to current analysis methods (APHA Standard Methods, 2005).
182 The chloride content was determined by the Mohr method, based on the formation of
183 silver chloride, an insoluble salt, detecting the turning point by the appearance of a red
184 precipitate of Ag_2CrO_4 , a compound used as an indicator (APHA Standard Methods,
185 2005). Sulfates were determined following the Rodier (1981) nephelometric technique.
186 The nitrate content is determined by second-derivative ultraviolet spectroscopy
187 following the Sempere et al. (1993) methodology.
188 The method for the determination of phosphorous is based on the formation of a
189 phosphomolybdic complex in an acid medium, reduced by ascorbic acid, producing a
190 blue coloration that is measured at 825 nm. The phosphorous is measured as a
191 phosphate ion.
192 The method for determining ammonium is based on the development of indophenol
193 blue by reaction of ammonium ions treated with a solution of sodium hypochlorite and
194 phenol in the presence of nitroprusside acting as a catalyst. The Na^+ , K^+ , Mg^{2+} , and Ca^{2+}
195 ions are measured directly in the sample or in appropriate dilutions by atomic emission
196 spectrophotometry in the case of the first two ions, and by atomic absorption for the last
197 two.

198 **2.7 Statistical Analysis**

199 Statistical analysis (based in the Student's t-test at 95% and ANOVA F test) were used
200 to determine the statistical significance of the treatments and differences between
201 means. Simple linear regression analysis was applied to the developed experimental
202 data. The squared correlation coefficient (R^2) represents the proportion of the variation
203 of a variable that is explained by its linear association with another variable.

204

205

206

207 **3 Results and discussion**

208 **3.1 Substratum and sewage sludge properties**

209 The substratum used has an alkaline reaction indicating that most nutrients could
210 manifest problems of availability (Jordán et al., 2004). In cases like this, acidifying
211 amendments are necessary to lower the pH, facilitate element mobility, and improve the
212 soil structure. The substratum has a relatively low nutrient content. Calcium is the
213 element found in the greatest proportion in the waste materials. Potassium and Na
214 concentrations (Table 1) are similar to the surrounding degraded agricultural soils
215 (Jordán et al., 2004). This is reflected in the soil's electrical conductivity because these
216 cations, especially Na, have a high mobility (Jordán et al., 2004; Jordán et al., 2008).
217 The equivalent calcium carbonate content is very high, as it is typical for these types of
218 residues. The organic matter content is very low (Table 1), just like that for available
219 phosphorous and Kjeldahl nitrogen compared with the desired normal content for a
220 cultivated soil.

221 The biosolid (Table 2) presents low contents of P and K, with medium contents for
222 calcium and magnesium, all within the ranges cited by Juárez et al., (1987). The total
223 Na content is of some importance, but cannot be considered dangerous for the soil
224 (Jordán et al., 2004). Analytically control of the sludge at the time of its incorporation is
225 important, especially with regards to Na, as this element may cause soil salinity
226 problems and alter its structure (Moreno Sánchez et al., 1986). The C/N ratio is 12,
227 indicating that the organic matter is partially mineralized and, therefore, the sludge can
228 enhance soil fertility (Hernández Fernández et al., 1986; Roldán et al., 1994). The
229 sewage sludge selected has an organic matter content that is very suitable for
230 agricultural use (Table 2).

231

232 **3.2 Leachate analysis**

233 The leachates of the mineral substratum can serve as a point of reference for possible
234 contamination that may appear in groundwater when sewage sludge is applied as an
235 amendment (Jordán et al., 2008). There are some physical–chemical parameters
236 concerning water that is destined to spill into aquatic resources. Some physical-
237 chemical parameters concerning water that can attain the aquatic resources are regulated
238 by the European Union Directive 76/464, transposed to the Spanish norm by the
239 Hydraulic Public Domain Regulation, approved by Royal Decree 849/1986. The Public
240 Administration of Water Regulation and the Hydrologic Plan, approved by RD
241 927/1988, states that the water quality objectives will be defined in the respective
242 Hydrologic plans, depending upon the foreseen water uses. The values in Table 5 are
243 solely illustrative, since in the case of leachates (washing waters) it does not deal with
244 waters directly spilled into the aquatic resource. Furthermore, the inexistence of a limit
245 value for nitrates calls attention.

246

247 **3.2.1 pH**

248 The pH values were found to be within the legislation limit value range (5.5-9). No
249 significant changes in the pH were produced between treatments (7.24-8.38); an
250 acidifying trend was only seen in the first and third sampling of the treatments. Over
251 time, it was observed that the pH values were more similar to that of the irrigating water
252 (7-7.8) before being added to the soil. The lowest pH values coincided with the
253 beginning of the experiment (incorporation of residual matter and beginning of
254 irrigation) and when the greatest degradation of the organic matter appears to have
255 occurred, between the second and third irrigation. Similar results have been obtained by

256 other authors using quarry spoils (Alcañiz et al., 1997; Almendro-Candel et al., 2007;
257 Jordán et al., 2008).

258 **3.2.2 Electrical conductivity**

259 With respect to the control ($< 800 \mu\text{S/cm}$), an increase in electrical conductivity was
260 observed in the water collected from the columns treated with biosolid (Table 5). This is
261 due to the resulting washing of the soluble salts contained in the biosolid applied to the
262 soil. The electrical conductivity values were closely related with the dose of biosolid
263 applied. Thus, the electrical conductivity values were only worrisome during the first
264 three irrigation applications ($6700\text{-}1234 \mu\text{S/cm}$); beginning with the fourth and
265 particularly the fifth ones, the electrical conductivity values stabilized as the irrigations
266 evidently washed out the salts ($< 500 \mu\text{S/cm}$). The electrical conductivity limit value is
267 $< 1000 \mu\text{S/cm}$. These values will be met from the fourth irrigation onward, while the
268 values up to that point were far superior. However, the quality of the aquifer's
269 groundwater is quite poor, reaching conductivities of $5000 \mu\text{S/cm}$, and so EC would not
270 represent any environmental risk to the aquifer.

271

272 **3.2.3 Inorganic nitrogen forms**

273 Two of the three inorganic forms of nitrogen in the leachates are discussed: nitrates and
274 ammonium (Table 5). The nitrites analyzed in the wash water were close to the
275 detection limit of the technique used. Their results are not discussed because they were
276 not significant.

277

278 An increase in nitrate concentration was observed in the water resulting from the wastes
279 treated with sludge with respect to the control waste (Table 5). The treatments with high
280 biosolid doses ($90,000 \text{ kg/ha}$) were those that contributed to the higher nitrate contents

281 to the water. The highest NO_3^- concentration in the leachates occurred in irrigations 1,
282 2, and 3 (2510-268 mg/L). From the fourth irrigation onward, the leaching of this anion
283 was scarce (< 100 mg/L). The nitrates exceeded the recommended values in the two
284 treatments. In any case, these high nitrate concentrations would drop with the
285 restoration and development of vegetative cover, which would assimilate a large portion
286 of the nitrates. Significant nitrate concentrations appeared that may pose an
287 environmental contamination risk. The Code of Good Agriculture Practices does
288 recommended the specific quantities of nitrogen to applied per hectare annually, which
289 in the case of manure, oscillates between 170 and 210 Kg/ha. However, there is no
290 mention of biosolid. Similar experiments carried out by Almendro Candel et al. (2007)
291 demonstrated that this mineral column, under the conditions prevailing in this study,
292 does not retain NO_3^- .

293 Nitrification is favored with good aeration and free drainage (Skiba and Ball, 2002).
294 This effect increased in the combined treatment of sewage sludge compost and saline
295 wastes. For both treatments with sewage sludge compost, nitrate leaching over time
296 could be associated with the biological activity of the biosolids, and the evolution
297 followed a similar trend to that observed by Santibáñez et al. (2007) in Chilean mining
298 tailings.

299
300 The values of ammonium were only significant for the first irrigation (0.2-1.6 mg/L).
301 This cation increased with the biosolide dose, and the differences decreased over time.
302 The ammonium quantities were far inferior to those obtained for NO_3^- .

303

304 ***3.2.4 Anions***

305 The limit values for the chlorides (< 700 mg/L), phosphates (< 27 mg/L), and sulfates

306 (< 800 mg/L) were very superior to those obtained in all the irrigations. A certain
307 tendency of the phosphorus to increase with the sludge dose treatment was observed.
308 The highest values (< 17 mg/L) were obtained in the columns filled with stripped soil
309 (D) and with the applications equivalent to 90,000 kg/ha of composted and treated
310 sludge. The recommended concentration limits (Table 5) were not exceeded in any
311 treatments. High natural limestone in the waste impeded in part the displacement and
312 loss of the P contained in the sewage sludge, which can precipitate as calcium
313 phosphate (Albiach et al, 2001).

314 Chlorides and sulfates are involved in plant nutrition (Almendro et al., 2014).
315 Furthermore, they are very relevant quality control parameters for water. Significant
316 differences appeared in the chlorides between the treatments with biosolids and control.
317 Probably, the most influential factor when determining the Cl⁻ in the leachates is the
318 contribution from the sewage sludge, as the substrata used (Z and D) contains abundant
319 salts. In fact, in the first two irrigations, high chloride values resulted in the control
320 columns with the presence of limestone spoils (Z or Z+D) with lower values in the
321 control columns filled with stripped soil (D). This observation demonstrates the
322 contribution of chlorides by the washing of the limestone spoils (Z) used as a mineral
323 substratum. The chlorides were completely washed out in the first three irrigations. In
324 the case of sulfates the highest values were reached in the second irrigation application.
325 This circumstance supports the fact that the organic sulfur may have undergone organic
326 matter mineralization processes and appeared in the leachates most significantly in the
327 third sampling.

328

329 **3.2.5 Cations**

330 From the environmental point of view, the concentrations of Ca, Mg, Na and K in the

331 leachates pose no risk (Nogués et al., 2000). The concentration of soluble K in the
332 sludge does not appear to produce an increase of this element in the leachates. It is
333 possible that the clayey nature of this substratum limits the displacement and loss of this
334 nutrient, which can be relatively adsorbed by the clay minerals (Pond et al., 2005). For
335 Na, a clear concentration increase was noticed in the first and second irrigation with the
336 treatment that was not significant for the remaining samplings. The concentration of Ca
337 seems to increase with the treatment and the sludge dose applied. Over time, its
338 tendency is to decrease. The soil reaction with the sludge appears to have increased the
339 concentration of soluble Ca, as it appeared in the leachates in considerable
340 concentrations. The Mg increased significantly with the treatment, and diminished over
341 time.

342

343 **3.3 Correlations**

344 The electrical conductivity correlated well with the concentrations of the cations
345 analyzed in the leachates, with the exception of K^+ (Table 6). In the case of Ca, this
346 correlation was excellent. It is obvious that the Ca in the leachates comes from both the
347 substratum used and the sludge applied as organic amendment.

348 Electrical conductivity correlations with the anions concentrations presented a
349 heterogeneous behavior. This correlation was excellent with NO_3^- , but less so for either
350 sulfate or chloride, as it could be expected (Table 6). This may be due to greater
351 mobility and concentration of the nitrate (Almendro-Candel et al., 2007). The resulting
352 nitrate values in the first irrigation applications were very high and, consequently, they
353 were washed out quickly. In the case of chlorides, the concentrations were lower.

354 Sulfates are salts that have a lower solubility than halite. There were significant
355 correlations between sulfate and Mg, Ca, and K. The highest correlation was obtained

356 with Mg^{2+} that comes mainly from the substratum formed by magnesium limestone and
357 dolomite subjected to a crushing process in the quarry plant. This may be due to the
358 presence of the epsomite ($MgSO_4 \cdot 7H_2O$) having a higher solubility (Jordán et al., 2004;
359 Jordán et al., 2008) than gypsum ($CaSO_4 \cdot 2H_2O$).

360 The good correlations obtained between the nitrate and the alkaline earth elements and
361 Na^+ were mainly due to their rapid percolation through the dissymmetrical columns,
362 above all in the first irrigation (Table 7). Chlorides could replace nitrate and nitrite, and
363 ammonium could also be exchanged by sodium, in the surface and exchange complex
364 of the residues (Santibáñez et al., 2007). After the first irrigations an equilibrium in
365 these materials is expected and the biological activation of the media may determine the
366 increment or descent of nitrogen in the leachates (Santibáñez et al., 2007).

367

368 **4 Conclusions**

369 The good correlations obtained for some physical-chemical parameters can help to
370 establish indicators of environmental quality of leachates over time. Therefore, constant
371 monitoring of water quality by selecting the most appropriate indicators is
372 recommended. As for the environmental risk with respect to the contamination of
373 aquifers, significant nitrate concentrations appeared that may pose an environmental
374 contamination risk. Irrigation scheduling should be an important part of a management
375 plan in limestone quarries reclamation.

376

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378

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FIGURE CAPTIONS

485 **Figure 1.** Columns used in the experiment.

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