

1           **Perspectives and challenges associated with the determination of new**  
2           **psychoactive substances in urine and wastewater – A Tutorial**

3

4 L. Bijlsma <sup>a\*</sup>, R. Bade<sup>b\*</sup>, F. Been<sup>c</sup>, A. Celma<sup>a</sup>, S. Castiglioni<sup>d</sup>

5

6 <sup>a</sup> Environmental and Public Health Analytical Chemistry, Research Institute for Pesticides and Water,  
7 University Jaume I, 12071 Castellón, Spain

8 <sup>b</sup> University of South Australia, UniSA: Clinical and Health Sciences, Health and Biomedical Innovation,  
9 South Australia 5000, Australia

10 <sup>c</sup> KWR Water Research Institute, Chemical Water Quality and Health, 3430 BB Nieuwegein, the  
11 Netherlands

12 <sup>d</sup> Istituto di Ricerche Farmacologiche Mario Negri - IRCCS, Department of Environmental Health  
13 Sciences, 20156 Milan, Italy.

14

15 \*Corresponding authors:

16 Lubertus Bijlsma (ORCID: 0000-0001-7005-8775), Environmental and Public Health Analytical  
17 Chemistry, Research Institute for Pesticides and Water, University Jaume I, Avda Sos Baynat s/n,  
18 12071 Castellón, Spain. E-mail address: [bijlsma@uji.es](mailto:bijlsma@uji.es)

19 Richard Bade (ORCID: 0000-0003-2724-9183), University of South Australia, UniSA: Clinical and Health  
20 Sciences, Health and Biomedical Innovation, South Australia 5000, Australia. E-mail address:  
21 [Richard.Bade@unisa.edu.au](mailto:Richard.Bade@unisa.edu.au)

22 **Abstract**

23 New psychoactive substances (NPS), often designed as (legal) substitutes to conventional illicit drugs,  
24 are constantly emerging in the drug market and being commercialized in different ways and forms.  
25 Their use continues to cause public health problems and is therefore of major concern in many  
26 countries. Monitoring NPS use, however, is arduous and different sources of information are required  
27 to get more insight of the prevalence and diffusion of NPS use. The determination of NPS in pooled  
28 urine and wastewater has shown great potential, adding a different and complementary light on this  
29 issue. However, it also presents analytical challenges and limitations that must be taken into account  
30 such as the complexity of the matrices, the high sensitivity and selectivity required in the analytical  
31 methods as a consequence of the low analyte concentrations as well as the rapid transience of NPS  
32 on the drug market creating a scenario with constantly moving analytical targets. Analytical  
33 investigation of NPS in pooled urine and wastewater is based on liquid chromatography hyphenated  
34 to mass spectrometry and can follow different strategies: target, suspect and non-target analysis. This  
35 work aims to discuss the advantages and disadvantages of the different data acquisition workflows  
36 and data exploration approaches in mass spectrometry, but also pays attention to new developments  
37 such as ion mobility and the use of *in-silico* prediction tools to improve the identification capabilities  
38 in high-complex samples. This tutorial gives an insight into this emerging topic of current concern, and  
39 describes the experience gathered within different collaborations and projects supported by key  
40 research articles and illustrative practical examples.

41

42 **Keywords**

43 New psychoactive substances; biological samples; wastewater-based epidemiology; monitoring  
44 strategies; mass spectrometry; ion mobility separation

## 45 **1. Introduction**

46 New psychoactive substances (NPS) are continually evolving and introduced in different ways in the  
47 drug market. The NPS retail market is characterized by its dynamic nature and the large number of  
48 substances covering a broad range of drug categories [1,2]. Whereas most NPS disappear after a short  
49 time, others seem to establish a niche market [2,3]. They are often introduced as legal substitutes for  
50 known controlled drugs, but also explored for their novel effect. Some substances have been known  
51 for years and are now misused for recreational purposes, but most NPS are newly synthesized with  
52 little or no safety data regarding their short or long-term toxicity. Furthermore, purity and composition  
53 of products containing NPS are often not known, which places users at an even higher risk compared  
54 to well-known conventional illicit drugs [1,2]. The NPS market is extremely diverse and differs between  
55 countries. Governments have responded in different ways to the NPS market, but have not been able  
56 to act upon all the NPS which have emerged in an effective way in terms of penalizing its supply and  
57 use [4]. Hence, NPS continue to cause public health problems [5,6] and challenge healthcare  
58 professionals, toxicologists and policymakers in terms of identification, prevention, treatment and  
59 control.

60 The Early Warning Systems (EWS) established by the European Monitoring Centre for Drugs and Drug  
61 Addiction (EMCDDA), Europol and the United Nations Office of Drugs and Crime (UNODC) play a key  
62 role in collecting data on new NPS appearing on the market. This information together with indications  
63 of the health and social risks associated with these substances is pivotal to respond to the emergence  
64 of NPS [7]. Analytical chemistry has a prominent role in gathering more thorough data which allows  
65 to better understand the situation of NPS use in the population. To complement the existing sources  
66 of information and improve our knowledge about the categories and characteristics of NPS present  
67 on the market, the application of appropriate analytical strategies is of utmost importance.

68 The discovery and characterization of new substances in commercially available products and drug  
69 seizures is an important source of information for EWS. Since reference standards for unambiguous  
70 confirmation of the identity are often not available, a combination of several techniques, such as  
71 nuclear magnetic resonance (NMR), liquid chromatography (LC) coupled to high resolution mass  
72 spectrometry (HRMS), gas chromatography mass spectrometry (GC-MS) and X-ray crystallography, is  
73 normally applied [8–11]. Although there is a correlation, the identification of new substances in seized  
74 products mainly gives information on the NPS available on the market rather than information on the  
75 prevalence of use. Therefore, the analysis of biological samples is needed, but this implies a different  
76 analytical strategy to deal with the complexity of the matrix and the low analyte concentrations  
77 normally present in the samples [12–14].

78 The analysis of biological samples can be considered a frontline in the detection of consumed NPS.  
79 Samples of individuals can be collected from, for example, hospital emergency rooms, drug testing  
80 campaigns or post-mortem examinations, where concentrations of some NPS in acute intoxications  
81 may be relatively high. This may facilitate the identification of hitherto unknown intoxicants by means  
82 of the abovementioned analytical techniques [3]. However, it does not give a full picture of NPS use  
83 within a community, rather individuals, and the analyses of many samples required to have a wider  
84 picture is time consuming and expensive. In contrast, pooled urine and urban wastewater can  
85 anonymously provide information of many people in one single aggregated sample. Although the  
86 dilution factor can be rather high in these matrices, for example dilution of the sample with urine of  
87 non-consumers or water used in households and industry, it has demonstrated its utility for  
88 community-wide monitoring of illicit drug use and showed possibilities for getting complementary  
89 insight into the consumption and diffusion of NPS use [15–20].

90 Liquid chromatography hyphenated to tandem mass spectrometry instruments (LC-MS/MS) with  
91 triple quadrupole mass analyzers (QqQ) or hybrid HRMS/MS systems are the preferred analytical  
92 techniques that have the required high sensitivity and selectivity to deal with the challenges related  
93 to the screening of NPS in pooled urine and wastewater. Furthermore, the polar characteristics of  
94 most NPS and their metabolites, as well as the sample matrix, make them compatible with these  
95 techniques. This article aims to discuss the advantages and disadvantages of relevant mass  
96 spectrometry (MS) data acquisition workflows and data exploration approaches to confront the low  
97 analyte concentrations and ever-changing NPS market and will be supported using key research  
98 articles and illustrative practical examples. This tutorial is not intended to be an extensive review of  
99 the existing literature, but to give an insight into this timely topic and describes the experience  
100 gathered within different collaborations and projects. It also pays attention to new developments such  
101 as ion mobility separation (IMS) and the use of *in-silico* prediction tools to improve the identification  
102 capabilities.

103 **2. Sample collection and sample treatment**

104 Well-designed protocols for sample collection and storage, and versatile sample treatment of pooled  
105 urine and wastewater are essential for getting data that provide meaningful information on NPS use.  
106 The collection of anonymous pooled urine samples from portable street urinals has recently  
107 demonstrated its utility to detect the use of recreational drugs, including NPS [18,19]. Generally,  
108 multiple samples are taken from various urine reservoirs, over a 12-hour period, and then mixed to  
109 form pooled urine samples. This sampling method can be applied in cities where stand-alone urinals  
110 are routinely used at weekends [19], but can also be used for monitoring specific night time settings  
111 or recreational events such as music festivals [20–24]. Sampling urine aliquots from urinals ensures  
112 the collection of anonymous and representative samples and results may reflect the direct use of NPS.  
113 Yet, some limitations are related to the fact that urinals are designed for male use only and normally  
114 have no ‘flushing’ mechanism [25]. Thus, the number of contributors to the samples is unknown and,  
115 although quantitative analysis is possible, the comparison of concentrations gives little additional  
116 insight rather than a qualitative overview of the actual use of a certain drug compared to the other  
117 substances quantified in that specific sample.

118 Wastewater analysis may circumvent these limitations by providing anonymous population-  
119 normalized information of an entire community and has recently been explored to gather information  
120 on NPS use [3,15,17,26,27]. The successful application of wastewater-based epidemiology for  
121 assessing spatial differences and temporal changes in illicit drug use has been demonstrated [28,29]  
122 where population-normalized data can be calculated taking into account the measured concentration,  
123 the daily flow rate of sewage and the number of people connected to the wastewater treatment plant  
124 (WWTP) [28,30]. Specific sampling protocols have been developed to obtain representative 24-hour  
125 composite wastewater samples collected at the inlet of a WWTP [28]. In addition, a standardized  
126 questionnaire facilitates the collection of relevant meta-data such as the daily flow rate of sewage and  
127 the number of people connected to the WWTP [31]. This meta-data allows quantitative population-  
128 normalized information for a limited number of target NPS to be explored. The information provided  
129 by wastewater analysis can be integrated with existing epidemiological data because of the unique  
130 ability to provide objective, updated and nearly real-time information on drug use [16,32].

131 One sampling technique not yet fully explored but with potential for monitoring NPS in wastewater is  
132 passive sampling [33], which ensures the concentration of analytes from longer periods (days or  
133 weeks) and increases the possibility to detect substances with low prevalence of use. The main  
134 advantage is that passive samplers, consisting of polymeric-based sorbent material, deployed for  
135 longer periods, can accumulate trace analytes on the sorbent during this period. Moreover, as some  
136 NPS might be consumed sporadically (and thus might not always be present in wastewater), one does

137 not need to collect multiple wastewater samples, which all eventually need to be processed *i.e.*  
138 increasing labor costs. Hence, this technology offers practical and economic advantages for gathering  
139 long-term data. But it has also some challenges related to calibration and quantification, since they  
140 require knowledge about uptake and diffusion of the different substances and are subject to the  
141 variability associated with NPS stability and environmental factors (e.g., flow rates, biofouling) [33,34].  
142 The uptake of target analytes on sorbent materials needs, therefore, to be determined prior to  
143 deployment in the sampling site.

144 Stability of NPS is an important aspect of sample collection for both pooled urine and wastewater  
145 analysis. While specific stability studies in pooled urine samples are lacking, they have been carried  
146 out on urine samples for forensic toxicology purposes. Metabolites of synthetic cannabinoids have  
147 been shown to be stable up to 14 days when refrigerated [35]. Many synthetic cathinones,  
148 benzodiazepines and amphetamine-type derivatives are very stable under freezing (-20 °C) storage  
149 conditions for months-years. However, when stored at room temperature or even refrigerated,  
150 degradation of these compounds can occur within days [36–38]. Therefore, it is recommended to  
151 freeze pooled urine samples immediately upon collection to avoid degradation. Regarding  
152 wastewater, it has been shown that acidification to pH 2 improves the stability in both filtered and  
153 unfiltered wastewater for up to 14 days for a wide variety of NPS such as cathinones,  
154 phenethylamines, opioid-derivatives and amphetamine-like stimulants [39]. If samples cannot be  
155 acidified, it is recommended that they are kept either refrigerated (4 °C) or frozen (-20 °C) for no longer  
156 than one week prior to sample processing [39–42]. Several synthetic cannabinoids have been shown  
157 to be unstable at pH 2 and in raw wastewater *i.e.* the hydroxypentyl metabolites of JWH 122, AM  
158 2201, RCS-4 and JWH 073, while JWH 018 n-pentanoic acid, JWH 073 N-butanoic acid and JWH018 N-  
159 5-hydroxypentyl were stable at room temperature for up to 24 hours [42]. Moreover, the use of  
160 sodium metabisulfite as a preservative has been recommended to improve the stability of synthetic  
161 cannabinoids [43].

162 A non-selective and versatile sample preparation protocol for the enrichment and clean-up of samples  
163 capable of retaining a wide range of NPS with broad physicochemical properties is preferred and  
164 applied by the vast majority of reported studies. Pooled urine samples are usually treated by  
165 performing a hydrolysis step to cleave drug-glucuronide conjugates with  $\beta$ -glucuronidase and  
166 arylsulfatase prior to solid-phase extraction (SPE), liquid-liquid extraction and/or dilute and shoot  
167 techniques [21,44,45], while wastewater samples do generally not require this hydrolysis step due to  
168 in-sewer deconjugation [46–50] and are normally filtered and solid-phase extracted [17], although a  
169 less labor-intensive and quicker preparation procedure following the QuEChERS principle has also  
170 been applied [51]. In order to cover the broadest range of substances possible, multiple SPE cartridges

171 or cartridges consisting of several layers with different stationary phase chemistries can be used  
172 [27,52]. The use of more cartridges implies several separate extractions, yet these can be optimized  
173 to specific NPS categories of interest such as cathinones or synthetic cannabinoids [15,21]. Typically,  
174 cartridges containing polymeric-based SPE sorbents with reversed phase (RP) properties built of  
175 generic hydrophilic and lipophilic balanced monomers or strong cation-exchange mixed mode  
176 sorbents incorporating RP copolymers are used. For the latter, samples should be acidified to pH 2-3  
177 to ensure that the analytes are positively charged during extraction [53]. This especially aids the  
178 recovery of cathinones, amphetamine-like stimulants, opioid derivatives and phenethylamines  
179 [21,39,45,54,55]. Online SPE has also been utilized for a limited number of NPS using a RP cartridge,  
180 with satisfactory recovery (*i.e.* 70-120%) [56]. LLE has been shown to aid in the detection of synthetic  
181 cannabinoids in pooled urine [57] and wastewater [43,58]. For wastewater studies, it is important to  
182 note that the removal of the solid fraction through filtration can greatly affect the overall recovery of  
183 synthetic cannabinoids due to their lipophilicity. Therefore, when performing wastewater analysis,  
184 both the aqueous and particulate fraction should be extracted together for optimal recovery of  
185 cannabinoids.

186 Although both pooled urine and wastewater analyses incorporate SPE, there is a much lower pre-  
187 concentration factor needed for pooled urine, with initial volumes of 1-2 mL, due to the generally  
188 higher concentrations found [23,45,57]. Furthermore, lower pre-concentration results in less matrix  
189 effects and potentially an improved chromatographic performance. Higher pre-concentration factors  
190 in wastewater are commonly applied to deal with the very low concentrations of NPS expected in  
191 these samples. However, this can also result in strong matrix effects due to the pre-concentration of  
192 unremoved components present in the sample extract. Matrix effects are alterations of the MS signal  
193 (enhancement or suppression), which have been linked to co-eluting interferences such as proteins,  
194 lipids, sugars or salts, that affect the ionization process [59]. Frequently, isotopically labelled internal  
195 standards (ILIS) are used as surrogates and added to samples prior to processing (*i.e.*, SPE) or analysis  
196 (in the case of dilute and shoot approaches applied in pooled urine analysis), to account for potential  
197 matrix effects, but also to correct for potential errors due to sample preparation. Ideally ILIS of the  
198 corresponding NPS are used as they are supposed to be affected in a similar manner as their non-  
199 labelled counterparts. However, ILIS are often expensive and not always commercially available,  
200 especially in the case of NPS. Therefore, ILIS are regularly used to correct for several compounds  
201 [15,40]. Nevertheless, the performance of each ILIS for correcting matrix effects need to be carefully  
202 evaluated. When appropriate ILIS are unavailable, matrix effects may be minimized by applying an  
203 additional clean-up step, but also lower pre-concentration factors may occasionally be desired for  
204 some substances in order to reduce ionization suppression and increase their detection limit [27,60].

205 In general, even when ILIS are available, a reduction of matrix effects is recommended for better  
206 precision, sensitivity and robustness in complex matrix samples [60].



### 207 3. Chromatographic separation

208 Good chromatographic separation is important to reach the required levels of selectivity, sensitivity  
209 and identification power to monitor NPS through wastewater and pooled urine analysis. GC-MS has  
210 been applied for the determination of NPS in urine. However, because of the high levels of selectivity  
211 and sensitivity provided by this technique, it requires the derivatization of the analytes which results  
212 in a more time-consuming and less generic sample treatment [61,62]. Alternatively, LC-MS allows the  
213 determination of compounds with a broad range of polarity, low volatility and thermolability with the  
214 application of more generic sample treatment strategies. In addition, the aqueous nature of the  
215 matrices makes LC-MS fully compatible with the determination of NPS in wastewater and pooled urine  
216 samples [63].

217 Reverse-phase LC (RPLC) separates compounds within the range of low-polarity to non-polarity.  
218 Therefore, it seems to be the most suitable chromatographic technique to achieve generic and good  
219 chromatographic separation especially for wide-scope monitoring of NPS. Consequently, the vast  
220 majority of studies dealing with multi-residue methods in wastewater and/or pooled urine samples  
221 applied RPLC as the separation technique [13,23,25,64,65]. However, more polar (or ionic) substances  
222 such as amphetamine-like stimulants or synthetic cathinones and their metabolites, might require  
223 more specific methodologies. Recent developments in column chemistries and improvement in  
224 robustness of existing stationary phases allowed the analysis of more particular scenarios. Hydrophilic  
225 interaction LC (HILIC) is an alternative approach to effectively separate small and highly polar NPS. For  
226 example, Kinyua et al. [55] successfully developed a multi-residue methodology for the determination  
227 of 7 synthetic cathinones and amphetamine-like stimulants by means of HILIC separation.

228 Additionally, enantiomeric analysis has also been explored for the determination of NPS [66–68].  
229 Chiral NPS are usually consumed as racemic mixtures of different forms (*i.e.* with an enantiomeric  
230 fraction (EF) between the two forms of approximately 0.5), even though both forms might differ  
231 quantitatively and qualitatively in the pharmacological activity [69]. Therefore, enrichment of the R  
232 (or S) form, depending on the stereoselective metabolism in humans, is expected in biological samples  
233 [66]. Consequently, an EF found in wastewater or pooled urine samples deviated from the original EF  
234 value could help in distinguishing between human consumption and direct disposal of unused  
235 substances [66]. Other chromatographic techniques such as capillary chromatography and  
236 supercritical fluid chromatography (SFC) are promising strategies for the monitoring of NPS. The  
237 improvement in sensitivity provided by capillary chromatography, especially for the small  
238 amphetamine-like structures, revealed a technique to explore for this purpose [26]. Also, recent  
239 developments in commercially available instruments has seen an increase in applications of ultra-high  
240 performance (UHP) SFC - MS/MS, in particular using (sub)supercritical carbon dioxide (CO<sub>2</sub>) with

241 various organic additives as mobile phase [58,70]. One of the main advantages of UHPSFC compared  
242 to conventional UHPLC is its increased chromatographic efficiency and resolution [71] also permitting  
243 the separation of several NPS isomers with good results [72].

#### 244 **4. Quantitative target monitoring**

245 As discussed above, the determination of NPS can be challenging due to the large number of  
246 potentially relevant compounds and the low concentrations expected in samples, in particular when  
247 considering wastewater and pooled urine. In fact, due to the often low prevalence of use of individual  
248 compounds, concentrations of these substances are often orders of magnitudes lower compared to  
249 conventional illicit drugs ( $< 10 \text{ ng L}^{-1}$ ) [55]. For this reason, targeted methods, specifically using LC-  
250 MS/MS with QqQ or ion-trap mass analyzers, have been implemented for the reliable identification  
251 and quantification of selected NPS in urine and wastewater samples [15,17,73,74]. The development  
252 of such quantitative target methods, however, requires access to reference standards for precursor-  
253 product ion transition selection in the Selected Reaction Monitoring (SRM) mode and MS parameters  
254 optimization. Identification and confirmation is achieved through the acquisition of at least two SRM  
255 transitions and matching of the retention time (RT) and ion-intensity ratios between the sample and  
256 reference standard [75,76]. The most sensitive SRM transition is commonly selected for the  
257 quantification at low concentration levels, whereas the second transition allows confident  
258 confirmation [26,40,46]. However, since NPS often retain high structural similarity, the risk of selecting  
259 common transitions is present and therefore the acquisition of more transitions (if feasible) is  
260 recommended to gain more confidence to the confirmation process. Hence, it is also important to  
261 understand fragmentation of each NPS as it allows the selection of specific product ions and avoid  
262 non-specific transitions such as a neutral loss of water or  $\text{CO}_2$ [77]. The latter is especially relevant to  
263 minimize potential matrix interferences when analyzing NPS at low concentrations in highly complex  
264 matrices such as pooled urine and raw wastewater samples. Although quantitative target monitoring  
265 can be performed using LC-HRMS instruments, their application in the field is limited due to the  
266 generally lower sensitivity compared to low resolution MS/MS instruments [17]. Hence, the advantage  
267 of low-resolution instruments for quantitative analysis lies in the robustness, selectivity and sensitivity  
268 which can be achieved by monitoring these specific precursor-product ion transitions. Combined with  
269 their high scanning speed, these instruments can monitor many transitions almost simultaneously,  
270 and consequently high-throughput, multi-residue methods that include many targeted NPS  
271 biomarkers, can relatively easily be developed.

272 Synthetic cathinones, phenethylamines, tryptamines and piperazine-derivatives have been  
273 quantitatively determined in pooled urine samples collected during weekends at specific night settings  
274 [25] or at music festivals [23]. Although data obtained from quantitative determination of NPS in  
275 pooled urine samples only gives an indication on the extent of use for an NPS compared to other  
276 substances found in a specific sample [23], these findings are still very valuable, as the application of  
277 these selective and sensitive target quantitative methods give high confidence and allows

278 confirmation of the NPS identified at low concentration levels. Synthetic cathinones are by far the  
279 most studied group of NPS in wastewater, followed by synthetic cannabinoids and phenethylamines.  
280 Studies using LC-MS/MS to monitor these substances have been carried out in Europe, Asia and  
281 Australia [15,17,26,39,40,78,79] and have shown spatial and temporal trends using population-  
282 normalized data. Although LC-MS/MS methods are highly sensitive and multi-residue methods can be  
283 developed, they have a major drawback, namely reference standard materials need to be available  
284 for method development as previously highlighted. Given the high number of NPS that have been  
285 detected in the market and their transient nature, reference standards are mostly available for only a  
286 limited number of compounds. Moreover, by the time reference standards become available, these  
287 compounds might have already disappeared from the market as they may have been less popular or  
288 added to the lists of regulated substances and can thus not be sold legally anymore. Further  
289 exacerbating the determination of these substances is the extent of their metabolism. There have  
290 been studies carried out on the metabolism of NPS using human liver microsome incubations to better  
291 understand the metabolism of certain NPS [80–84]. In addition, recent advances in computing power  
292 have permitted the development of comprehensive knowledge based software to predict the  
293 metabolic fate [85,86]. However, reference standards of most of the metabolites proposed are not  
294 commercially available and therefore unsuitable for quantitative target monitoring. Thus, quantitative  
295 target LC-MS/MS methods, although indispensable to achieve the highest sensitivity needed for  
296 certain types of substances (e.g., fentanyl and its derivatives), need to be complemented by other  
297 analytical approaches which allow a quick and broader monitoring, without the necessity for reference  
298 standards. Although low-resolution mass spectrometry (LRMS), especially tandem MS instruments,  
299 are highly appreciated in quantitative analysis, its application to qualitative analysis and capabilities in  
300 detecting unknowns is, limited due to the relative low resolving power (approximately 1 Da) and low  
301 sensitivity in full scan mode [77]. The use of HRMS offers new possibilities in the determination of NPS  
302 as well as circumventing some of the limitations of LRMS.

## 303 5. Qualitative screening approaches

304 HRMS presents strong potential for monitoring a large number of substances, due to its acquisition of  
305 accurate-mass full spectrum data at good sensitivity [63,77,87]. In order to facilitate the reading of  
306 this tutorial, terms that will be used in this section are defined below:

307 *Target screening* based on HRMS allows the qualitative screening of NPS after data acquisition based  
308 on large databases, thus evading the pre-selection of analytes for method development and the need  
309 of reference standards. However, the information included in the database is limited by the availability  
310 of reference standards. When reference standards are available, information such as accurate masses  
311 of fragment ions, adduct formation and RT can be included, whereas only the elemental composition,  
312 exact mass and theoretical isotopic pattern can be included when no reference standard is available.  
313 Although the acquisition of data is performed in an untargeted way, the approach is considered  
314 targeted and generally known as *suspect screening* [77,87], since the search is based on a list of target  
315 compounds that can be expected to be found in the samples. An advantage of this approach is that  
316 retrospective analysis can also be performed at any time from the acquired data to search for  
317 substances initially not considered and included in the database, such as novel NPS or newly  
318 discovered metabolites [88,89]. It should, however, be noted that the detection of some substances  
319 might be restricted by the sample treatment, the chromatographic conditions or the ionization  
320 efficiency [90], since usually a generic analysis is performed and no optimization has been executed  
321 for the NPS included in the database.

322 *Non-targeted screening*, without any selection of analytes, allows the investigation of any other NPS  
323 biomarker not included in the database. However, it implies an examination of each chromatographic  
324 peak and extensive investigation of its accurate mass spectrum. This process is challenging and time  
325 consuming and probably does not outweigh the rate of success in identifying of unknown NPS.  
326 Alternatively, the screening can be directed to discover related compounds of known NPS using  
327 characteristic mass spectral information and applying mass-defect filtering or common fragmentation  
328 pathways.

329 As a starting point for researchers interested in undertaking qualitative screening of NPS by HRMS,  
330 the review article written by Hernandez et al. [63] describing different mass spectrometric strategies  
331 for the investigation of illicit drug biomarkers in wastewater is recommended. Although similar  
332 strategies and identification criteria can be applied for the investigation of NPS in pooled urine and  
333 wastewater, the challenges are different due to the rapid turnover in the NPS drug market creating a  
334 scenario with constantly moving analytical targets and the often lower prevalence of use compared  
335 to conventional illicit drugs. Moreover, the structural similarities of NPS and their metabolites often  
336 requires increased identification confidence in order to minimize reporting false positives. In the text

337 below, practical examples are given to discuss different data acquisition workflows and data  
338 exploration approaches to illustrate how HRMS can help in the confident identification of NPS in high-  
339 complex pooled urine and wastewater samples.

340

#### 341 ***5.1. Acquisition modes for hybrid high resolution mass spectrometric systems***

342 The most commonly used HRMS analyzers are time-of-flight (TOF) and Orbitrap, which can be coupled  
343 with LC and possess high mass resolving power ( $> 20,000$  Full Width at Half Maximum (FWHM)) and  
344 mass accuracy ( $< 5$  ppm) for wide scope screening of NPS in pooled urine and wastewater [17,75,76].  
345 However, hybrid configurations, such as quadrupole-TOF (QTOF) or quadrupole-Orbitrap (Q-Orbitrap),  
346 are nowadays more the standard than the exception as they considerably increase the potential of  
347 HRMS for screening NPS [20,21,27,44,52,91]. When working in MS/MS mode, it is possible to record  
348 accurate mass product-ion spectra of previously detected candidates and obtain relevant structural  
349 information to allow suspected NPS to be confidently identified or disregarded as false positives.  
350 However, the simultaneous accurate-mass acquisition of both full-spectrum and product-ion spectra  
351 data is preferable and collects accurate mass data of both the (de)protonated molecules and its  
352 fragment ions in a single acquisition and without the selection of precursor ions.

353 In data-dependent acquisition (DDA) mode, the instrument first performs a “survey scan” from which  
354 the analyst chooses (or not) certain ions that fit specific criteria based on, for example, intensity  
355 thresholds. Ions for which these conditions are met, are then selected to be included in a list of  
356 preselected masses and fragmented to provide information-rich product ion scans. Unlike intensity  
357 thresholds, an inclusion (or exclusion) list allows large matrix interferences to be ignored, thereby  
358 facilitating the identification process and saving effort and time [27,52,63,92]. However, the size of  
359 the inclusion list (i.e., suspects to be fragmented) can adversely affect the cycle time of the instrument.  
360 Therefore, a decrease in the number of scans (or data points) across a chromatographic peak will  
361 occur, reducing its detectability. Moreover, any compound not included in the initial inclusion list  
362 cannot later be retrospectively analyzed, so the sample would have to be re-extracted and re-  
363 analyzed. Yet, there is a way around this limitation, utilizing complementary targeted and untargeted  
364 DDA. This technique initially conducts an MS scan followed by targeted MS/MS using an inclusion list  
365 and then untargeted MS/MS on  $n$ -selected precursors. For example, analysts can look at MS/MS of  
366 the  $n$  most abundant precursor ions, which would be of great utility for samples with high levels of  
367 NPS such as seizure samples [14,93]. However, the generally low concentration of NPS found in pooled  
368 urine and wastewater might mask the detection of low abundant peaks, and therefore, many NPS may  
369 remain undetected [94].

370 Data independent acquisition (DIA) allows the acquisition of accurate-mass full-scan spectra under  
371 different collision induced dissociation conditions within a single injection. This acquisition mode is  
372 known under different names depending on the manufacturer (e.g. All-ion-fragmentation (AIF), all-  
373 ion MS/MS, MS<sup>E</sup> and broadband collision-induced dissociation (bbCID)), where all ions generated in  
374 the ion source are sent to the collision cell for fragmentation without precursor ion selection or any  
375 predefined selection criteria. This alternation between full-scan and untargeted MS/MS events at low  
376 collision energy (LE) and high collision energy (HE), respectively, allows one to obtain information  
377 relating to the accurate masses of the (de)protonated molecule as well as their fragment ions.  
378 Furthermore, it conserves highly valuable information on adducts and isotopes since the quadrupole  
379 works as an ion guide [63,77]. The main limitation of DIA is that spectra are non-selective and contain  
380 product ions for all ions formed in the ion source. Hence, the interpretation can be challenging, since  
381 co-eluting compounds or matrix interferences may “contaminate” the spectra, and makes it difficult  
382 to associate product ions with the correct (de)protonated molecule [14,95,96].

383 Slightly different modes compared to the other DIA modes mentioned above in terms of specificity  
384 have been developed by manufacturers with the objective to have HE spectra approaching to MS/MS  
385 quality data. As an example, in Sequential Window Acquisition of all THEoretical fragment ion spectra  
386 (SWATH) mode, a TOF MS full scan at LE is acquired, alternated by SWATH experiments at HE obtaining  
387 MS/MS data by fragmenting only the (de)protonated molecules present in a much narrower window  
388 (e.g. 15 -25  $m/z$ ). In this way, SWATH can distinguish co-eluting compounds of different masses by  
389 having specific experimental mass fragmentation windows which filter out all masses not included in  
390 the specified mass range. This results in cleaner spectra, which facilitates identification [96,97]. This is  
391 a particular important point in the determination of NPS, which are notorious for the analytical  
392 challenges associated with common fragments. **Figure 1** shows the utility of SWATH in differentiating  
393 two co-eluting NPS, butyryl fentanyl with  $m/z$  351.2431 and furanylfentanyl with  $m/z$  375.2067 in a  
394 spiked wastewater sample. In the full scan acquisition at LE, it can be observed from the individual  
395 extraction ion chromatograms (XICs) that the two NPS seemingly elute at 12.50 min (**Figure 1A, top**),  
396 with the mass spectra at this RT showing both masses (**Figure 1A, bottom**). However, when applying  
397 SWATH, the HE experiments carried out at different mass windows ( $m/z$  340.2 – 357.4; **Figure 1B** and  
398  $m/z$  372.6 – 389.8; **Figure 1C**) allowed them to be distinguished by extracting the mass of each of these  
399 fentanyl derivatives in their corresponding acquisition window. With the mass of butyryl fentanyl and  
400 furanylfentanyl falling within separate experiments, they can be individually extracted and identified  
401 using cleaner spectra. This exemplifies the power of this acquisition mode in the elucidation of NPS.

402

403 [Insert **Figure 1** here: Identification of two co-eluting NPS, butyryl fentanyl ( $m/z$  351.2431) and  
404 furanylfentanyl ( $m/z$  375.2067) in a spiked wastewater sample using Sequential Window  
405 Acquisition of all THEoretical fragment ion spectra (SWATH). (A) overlapping extraction ion  
406 chromatograms (XICs) of the two NPS with chromatographic peaks eluting at 12.50 min (top);  
407 full scan acquisition mass spectra with low collision energy (LE) (10 V) at retention time  $12.50 \pm$   
408 0.10 min (bottom). (B) SWATH mass window  $m/z$  340.2-357.4, XIC at  $m/z$  351.24 (middle) and  
409 high collision energy (HE) mass spectra (bottom); (C) SWATH mass window  $m/z$  372.6-389.8, XIC  
410 at  $m/z$  375.21 (middle) and HE mass spectra (bottom)]

411

## 412 **5.2. Suspect screening**

413 Suspect screening approaches usually take advantage of home-made databases. However, the  
414 information included therein is limited by the availability of reference standards, as previously  
415 explained. When no reference standard is available, the minimum suggested requirements for a  
416 tentative identification is the accurate mass of the (de)protonated molecule and, at least, one  
417 significant fragment ion together with the corresponding isotopic pattern. This is in the line with  
418 proposed quality procedures recommended in other research fields [76,98]. The observed fragments  
419 need to be in accordance with the chemical structure and, preferably, in agreement with previously  
420 reported data in scientific literature or online spectral databases [27,52,99,100]. Ideally, reference  
421 standards are available, and information such as accurate masses of fragment ions, adduct formation  
422 and RT can be included, which allow unequivocal identification. However, this entails high costs due  
423 to the high number of compounds and, therefore, huge efforts have been devoted, in the recent years,  
424 to develop community-made or online mass spectral databases for NPS. The best known databases  
425 are NPS Data Hub [101] and HighResNPS [102,103] with more than 2800 and 3350 entries, respectively  
426 (date accessed: 26 June 2020). The HighResNPS library currently has active users from more than 10  
427 laboratories around the world with the intention to ensure up-to-date analytical information from the  
428 moment a specific NPS becomes available to a given participating laboratory [102]. These libraries are  
429 available to help and facilitate the screening of NPS and their metabolites [101,104–106].

430 In most laboratories, a suspect screening based on large home-made databases is often the first step  
431 for monitoring samples. Due to the high number of NPS and metabolites, the rapid transience of these  
432 compounds on the market, high costs and limited availability of reference standards, home-made  
433 databases are normally built of merely accurate masses of the (de)protonated NPS and fragment ions.  
434 Yet, the low concentration levels of NPS present in combination with strong matrix interferences  
435 makes the tentative identification of NPS challenging and remark often the necessity to perform some  
436 additional research or experiments to increase the confidence in the tentative identification. As an



437 example, **Figure 2** shows the tentative identification of 4-chloro- $\alpha$ -pyrrolidinopropiophenone (4-  
438 chloro- $\alpha$ -PPP) in a pooled urine sample. Its protonated molecule, the isotopic information related to  
439 the presence of one chlorine atom and at least one fragment ion was observed at accurate mass  
440 (**Figure 2A**). However, a known and abundant fragment of 4-chloro- $\alpha$ -PPP at  $m/z$  167.0258 [107]  
441 showed an undue high mass error (+143 ppm) under the initial screening conditions, which made the  
442 tentative identification of this NPS questionable. By increasing the mass resolution of the Orbitrap MS  
443 from 20.000 to 35.000 FWHM and zooming in the  $m/z$  range of the fragment, it was possible to  
444 distinguish three peaks at  $m/z$  167, one corresponding to the fragment ion  $m/z$  167.0258 (+5.3 ppm)  
445 of 4-chloro- $\alpha$ -PPP (**Figure 2B, bottom**). This allowed more confidence to be gained in the  
446 identification. Subsequently, the feature could be identified as 4-chloro- $\alpha$ -PPP by means of a reference  
447 standard. The latter is pivotal for the confirmation of the identity of the NPS. However, by using this  
448 approach, laboratories do not need to purchase all reference standards *a priori* to the analysis [108]  
449 and could prioritize those NPS for which more reliable evidence is obtained.

450

451 [Insert **Figure 2** here: Tentative identification of 4'-chloro- $\alpha$ -pyrrolidinopropiophenone (4-chloro- $\alpha$ -  
452 PPP) in a pooled urine sample. (A) Extracted ion chromatogram of 4-chloro- $\alpha$ -PPP and  $^{37}\text{Cl}$   
453 isotope (top); Product ion mass spectra of  $[\text{M}+\text{H}]^+$  at  $m/z$  238.10 (bottom). (B) Structure of 4-  
454 chloro- $\alpha$ -PPP (top); Zoom in the range of fragment ion with  $m/z$  167 at resolution (R) of  
455 35.000 Full Width at Half Maximum (FWHM) (bottom)]

456

457 Positional isomers or homologues are frequently the first choice to substitute banned NPS [109].  
458 Hence, NPS often have only minor modifications to a backbone structure and the structural similarities  
459 of NPS and their metabolites are often reflected by their common fragmentation pathways, this poses  
460 one of the principal challenges in suspect screening strategies. As an example, the analysis of a raw  
461 wastewater sample showed a chromatographic peak at 4.51 min giving a positive hit for the isomers  
462  $\alpha$ -methyltryptamine (AMT) and 5-(2-aminopropyl)indole (5-IT) based on the accurate mass of their  
463 protonated molecule and their fragment ions (**Figure 3A**). These two isomers share the same chemical  
464 backbone with the only difference being the position of the substituent (**Figure 3B and 3C, top**). The  
465 following MS fragment ions were found:  $m/z$  158.0954,  $m/z$  143.0724,  $m/z$  132.0799,  $m/z$  117.0577  
466 and  $m/z$  115.0541, with the most abundant fragment at  $m/z$  143.0724 (**Figure 3A, bottom**). The only  
467 difference, described in the literature, between the spectra of AMT and 5-IT resides in the relative  
468 intensities of the fragment ions [110]. The most intense fragment ion of 5-IT has an  $m/z$  of 130,  
469 whereas the most abundant fragment ion for AMT corresponds to  $m/z$  143. This slight difference in  
470 the fragmentation pattern (i.e. intensities) gave more confidence in the tentative identification of AMT

471 instead of 5-IT in this sample. Therefore, AMT was synthesized and a reference standard of 5-IT was  
472 donated by a collaborating laboratory. When comparing that empirical data to AMT and 5-IT reference  
473 standard MS fragment ions (**Figure 3b and 3c, bottom**), it can be observed that both substances share  
474 the same fragment ions (in nominal mass;  $m/z$  143,  $m/z$  130,  $m/z$  117 and  $m/z$  115) coinciding with  
475 the fragment ions observed in the sample, but that AMT indeed show a more abundant fragment ion  
476 with  $m/z$  143. This gave more confidence in the positive identification of this NPS and together with  
477 its RT, AMT could finally be confirmed.

478

479 [Insert **Figure 3** here: Identification of  $\alpha$ -methyltryptamine in a raw wastewater sample using QTOF  
480 MS. (A) feature detection of  $m/z$  175.1235 at 4.51 min (top, insert) together with the low  
481 collision energy (LE) spectra (top) and high collision energy (HE) spectra with emphasis on  $m/z$   
482 130-145 (grey areas) (bottom); (B) Structure, fragment ions, LE and HE spectra of  $\alpha$ -  
483 methyltryptamine; (C) Structure, fragment ions, LE and HE spectra of 5-(2-aminopropyl)indole]

484

### 485 **5.3. In-silico approaches**

486 In some cases, the instrument-specific parameters (i.e. accurate mass ions and isotopic patterns) do  
487 not suffice to tentatively propose a chemical structure, and, therefore, additional studies are required.  
488 For that purpose, predictive models have been used to filter out false positives and increase the  
489 confidence of compound identification when reference standards are unavailable or no information  
490 is within reach in previously reported data [27,111,112]. Aalizadeh et al. developed a RT prediction  
491 model using Quantitative Structure-Retention Relationships (QSSR) and Support Vector Machines  
492 (SVM) to model the RT data for both HILIC and RPLC with high accuracy [111]. A different approach  
493 was proposed by Bade et al. considering the application of Artificial Neural Networks (ANNs) for the  
494 development of a RT predictor for gradient-RPLC using a dataset of more than 500 compounds with  
495 an predictor accuracy of  $\pm 2$  min [112]. Such RT predictive tools are highly valuable for the  
496 determination of NPS in complex matrices as demonstrated by Diamanti et al. [27]. Since the  
497 availability of reference standards is limited, the suspect screening of NPS usually results in many  
498 candidate structures because of the structural similarity of many NPS, as for example, in the case of  
499 the two isomeric phenethylamines 3,4-methylenedioxy-N-hydroxyethylamphetamine (MDHOET) and  
500 N-hydroxy-N-methyl-3,4-ethylenedioxyamphetamine (EFLEA). The predicted RT using a QSSR  
501 predictor model matched the one for MDHOET and discarded the one for EFLEA, thereby reducing the  
502 number of candidates and increasing the confidence in the tentative identification of MDHOET in  
503 influent wastewater from Athens [27]. *In-silico* fragmentation tools, such as the MetFrag software, are  
504 pivotal in a suspect screening workflow. This software generates a predicted fragmentation of

505 molecules based on their structure and compare it to the empirical data gathered proposing a list of  
506 fitting candidates together with a scoring parameter [113,114]. However, it is common that many  
507 structurally related substances can be assigned to the empirical data with a similar score value [113],  
508 which is a drawback particularly for the investigation of NPS because of the similarity of several  
509 substances.

## 510 **6. Ion mobility separation coupled to high resolution mass spectrometry**

511 The recent development of the hyphenation of IMS with LC-QTOF MS instruments (LC-IMS-QTOF MS)  
512 represents an innovative tool for their application in target and non-targeted screening strategies. IMS  
513 separates ions depending on their size, shape and charge in a gas phase, (usually nitrogen or helium),  
514 and in the presence of an electric field [115]. Ion separation occurs in the millisecond time scale,  
515 making it compatible with fast TOF MS acquisitions [116]. The time an ion takes to travel through the  
516 mobility cell i.e. the drift time (DT), adds an extra dimension to the obtained chromatographic RT and  
517 accurate mass, which results in increased selectivity and improved identification, particularly in DIA  
518 modes [116,117]. The increased selectivity is translated into much cleaner and higher-quality spectra  
519 than conventional HRMS DIA spectra, since (de)protonated molecules and fragment ions of interest  
520 with the same DT can be aligned and separated from co-eluting matrix components. Although data  
521 sets inherently become more complex and more comprehensive, the utilization of IMS-HRMS  
522 instruments does not overcomplicate the data revision process thanks to the four-dimensional  
523 automatic feature detection. This allows the software to both deconvolute peaks based on  
524 chromatographic and MS data and align ions with the same RT and DT into unique features. Thus, LE  
525 and HE spectra are DT filtered for the deconvoluted ions (*i.e.* for each ion detected in the LE spectra  
526 its DT is used to correlate it with the fragment ions obtained in the HE spectra). Cleaner spectra can  
527 also be obtained by improving the chromatographic separation. Although improvements in the quality  
528 of the spectra often relies on spectral discrimination of the compounds, a good chromatographic  
529 separation is recommended especially when analyzing complex matrices such as pooled urine and  
530 wastewater that contain many co-eluting interferences. Yet, IMS provides an extra dimension of  
531 separation which fits between chromatography and MS and results in cleaner spectra, but without  
532 increasing the chromatographic run time or mass resolving power.

533 A further advantage of IMS is that Collision Cross Section (CCS) values can be derived from the DT and  
534 represent the surface of the sphere created by the ion when moving in the gas phase. Unlike DT, CCS  
535 is an instrument independent value, provided that the same drift gas and ion mobility calibration  
536 standards are used [116,118,119]. The importance of CCS values relies on the fact that they are robust  
537 across multiple platforms (*i.e.* deviation up to 2%), independent of the chromatographic conditions  
538 used and not affected by matrix composition [118–120]. CCS values depend on the calibration  
539 procedure applied, and the deviation between instruments is caused by the slight experimental  
540 variations in room temperature, gas pressures and other hardware settings. Hence, CCS is a parameter  
541 that can give support to MS-based compound identification in addition to RT,  $m/z$ , isotopic pattern  
542 and fragment ions. Finally, IMS enables, in theory, the separation of isomeric compounds not  
543 previously resolved using LC, since they are expected to have a different mobility in the drift cell, and

544 therefore different CCS values [121,122]. Although there is a relationship between the  $m/z$  and CCS,  
545 Bijlsma et al. [123] showed that a range of  $35 \text{ \AA}^2$  could be observed for molecules of approximately  
546 300 Da, therefore, demonstrating that no direct correlation between  $m/z$  and CCS could be established  
547 and that thus IMS may separate isomers.

548 **Figure 4** illustrates the benefits of IMS in terms of higher-quality spectra in DIA MS/MS events. In this  
549 example, a positive finding of ketamine in a wastewater sample is shown using an ion mobility  
550 separation QTOF MS (Vion from Waters). When searching for ketamine (with  $m/z$   $238.0993 \leq 5\text{ppm}$ )  
551 a chromatographic peak at a RT of 3.33 min was observed (**Figure 4A, top** (yellow arrow)). The  
552 corresponding conventional DIA MS<sup>E</sup> spectra (LE and HE) show many ions when no DT alignment is  
553 applied (**Figure 4B, top**) resulting in a base peak with  $m/z$  263.1386, which does not correspond to  
554 ketamine (i.e.  $m/z$  238.0993, highlighted in green). However, when applying the IMS MS<sup>E</sup> acquisition  
555 mode (HDMS<sup>E</sup>, High-Definition MS<sup>E</sup>), several co-eluting ions at 3.33 min are separated in the mobility  
556 cell, illustrated as red or black dots in **Figure 4A, bottom**. The DT of the ion with  $m/z$  238.0993 was  
557  $4.89 \pm 0.20$  ms and the corresponding fragment ions in this range, the blue highlighted areas, can be  
558 aligned. All other ions outside this area are filtered out, which results in much cleaner and easier to  
559 interpret spectra (**Figure 4B, bottom**). Despite the presence of some co-eluting interferences with  
560 similar DT, the resulting spectra contains fragment ions which could be primarily assigned to ketamine  
561 [124].

562

563 [Insert **Figure 4** here: Identification of ketamine in a wastewater sample using IMS QTOF MS. (A)  
564 feature detection of  $m/z$  238.0993 at 3.33 min and drift time (DT) 4.89 ms, yellow arrow (top);  
565 co-eluting ions at 3.33 min illustrated as red or black dots and separated by DT. Blue highlighted  
566 areas are the DT ranges of  $4.89 \pm 0.20$  ms at  $m/z$  238.0993 at low collision energy (LE) and high  
567 collision energy (HE) (bottom). (B) LE and HE mass spectra without IMS DT alignment (top); LE  
568 and HE mass spectra with IMS DT alignment (bottom)]

569

570 The additional cleaning of spectra provided by IMS is of particular relevance for the determination of  
571 NPS in challenging matrices such as wastewater or pooled urine where thousands of naturally  
572 occurring compounds can hamper the identification of these substances at the low concentration  
573 levels expected. Moreover, since the CCS value of a certain molecule is not affected by matrix  
574 composition, their utilization as an additional identification point in the determination of NPS pushes  
575 IMS-HRMS as a promising technique in the monitoring of these substances [125,126]. Therefore, the  
576 development of home-made or collaborative on-line databases including ion mobility data will  
577 enhance the efficiency of target NPS screening. However, as has been discussed earlier, due to the

578 lack of analytical standards for most of the NPS and metabolites and the still sparse accessibility to  
579 IMS-HRMS instruments in research centers, the availability of CCS values for these substances is still  
580 very limited. Hence, *in-silico* predictive tools similar to those for RT and MS fragmentation may help  
581 to increase the confidence in the identification of tentative candidates. Several data-driven CCS  
582 predictor systems have been developed for the prediction of CCS values for small molecules [123],  
583 pharmaceuticals and drugs of abuse [127] and metabolites [128]. As an example, the predictor  
584 reported by Bijlsma and Bade et al. [123] was developed using 205 CCS values for small molecules  
585 including pharmaceuticals, pesticides and drugs of abuse with ANNs for modelling the ion mobility  
586 data. Although the empirical variability of CCS measurements across instruments for a certain  
587 molecule is known to be up to 2%, with the developed CCS predictive model, the maximum deviation  
588 at the 95% confidence interval was only 6%. Mollerup et al. [127] were able to reduce the deviation  
589 in the predicted CCS to a 4%, consequently increasing the accuracy of the model. In the case of the  
590 predictor model developed by Zhou et al. [128], support vector regression was applied to the  
591 development of predictive models for different molecular adducts with median relative errors of  
592 approximately 3%. Regardless of the predictive model applied for the prediction of CCS, the utilization  
593 of these strategies facilitates the tentative identification of NPS in suspect screening strategies [125],  
594 especially when combined with RT and MS fragmentation predictive tools.

## 595 **7. Future perspectives**

596 The determination of NPS in pooled urine and urban wastewater has shown several challenges due to  
597 distinct factors as discussed in this manuscript. Current analytical instrumentation based on LC  
598 combined with LRMS and HRMS and the application of complementary data acquisition workflows  
599 and data exploration approaches helps to circumvent or confront certain barriers. However, more  
600 research related to NPS biomarkers is required and several trends in analytical chemistry, which is  
601 under continuous development, can be highlighted:

602

603 i. ***NPS biomarker selection***. The high number of existing NPS and the constant introduction of new  
604 compounds on the drug market creates a dynamic scenario of moving target biomarkers. Hence,  
605 monitoring of all NPS is complex and efforts could therefore be initially focused on NPS which are  
606 relatively high-dosed or frequently consumed and excreted (partly) unchanged such as  
607 amphetamine-like substances and cathinones. Especially since scant information on NPS  
608 pharmacokinetics is currently available, which complicates the choice of suitable biomarkers  
609 (parent substance or urinary metabolites) [129,130]. This is particularly relevant for synthetic  
610 cannabinoids and compounds like NBOMes that are highly metabolized in the human body  
611 [42,131,132] and for synthetic opioids that are consumed at very low doses [39], leading in both  
612 cases to very low concentration levels of the corresponding biomarkers in urine and,  
613 consequently, in wastewater. However, there are some published works on the metabolism of  
614 NPS [80–84] and different computational tools exist that predicts the metabolic fate of chemicals  
615 [86,87]. Although the proposed metabolites therein are generally not commercially available for  
616 quantitative target monitoring, these compounds should be included within screening databases  
617 as well as aiding in retrospective data analysis to ensure that the most appropriate analytical  
618 targets are investigated.

619 ii. ***Sample collection, storage and treatment*** of pooled urine and wastewater is pivotal for getting  
620 meaningful information on NPS use. Pooled urine analysis of samples collected from portable  
621 toilets and urinals give an informative snapshot of the NPS used, but is often limited to men only  
622 and it is difficult to extrapolate results to the total number of toilet users. All-gender toilets with  
623 an improved design, complying specific technical requirements like a flushing mechanism and a  
624 visitor counter could circumvent these limitations in future studies. Currently, daily composite  
625 wastewater samples are more representative and analysis provides population-normalized  
626 quantitative information on NPS. A best practice protocol to collect representative wastewater  
627 samples of an entire community is available [32] to ensure the comparability of results from  
628 different countries. However, wastewater is more diluted compared to pooled urine resulting in

629 lower concentrations, which may complicate the detection of some NPS. Passive sampling  
630 increases the possibility to detect substances with low prevalence of use, because of the sampling  
631 and concentration of analytes over a longer period of time. Yet, passive sampling also merely  
632 gives a snapshot and has several limitations that need to be overcome or optimized as previously  
633 described. Recent developments, using diffusive gradients in thin films which, in contrast to  
634 conventional samplers, consist of a diffusive and binding gel and are exposed to the medium, are  
635 less dependent to hydrodynamic condition (e.g. flow rates) and can hence overcome some of  
636 the limitations encountered with conventional passive samplers [133,134].

637 A relevant requirement for an NPS biomarker is its stability in pooled urine and wastewater in  
638 order to avoid any loss that can prevent detecting its use. Further work need to be addressed to  
639 test biomarkers stability and potential degradation or transformation in raw wastewater and  
640 urine [39–41,94]. Until more information is available, it is recommendable to store samples in the  
641 dark at -20 °C directly after sample collection in order to minimize possible degradation.

642 Sample treatment is very important to improve detection. However, a versatile sample treatment  
643 to retaining a wide range of NPS is not always feasible and specific treatments for certain NPS  
644 classes such as synthetic cannabinoids and synthetic opioids (i.e. high potency NPS such as  
645 fentanyl) need to be developed.

646 iii. Good **chromatographic separation** might seem less important when coupled to highly sensitive  
647 and selective mass spectrometers, although it can be essential in the detection and identification  
648 of NPS. Taking into account the many isomers or structurally related compounds and the often  
649 strong matrix effects, more effort could be put into chromatographic separation in future work.  
650 HILIC and enantiomeric analysis have demonstrated a strong potential to move a step forward  
651 into a more comprehensive determination of NPS in wastewater and pooled urine. Capillary  
652 chromatography and UHPSFC-MS/MS have also been explored. Yet, some concerns have also  
653 been raised related to the robustness of the technique to routinely analyze complex matrices.  
654 Future developments in terms of more robust column chemistries will open a new scenario for  
655 the monitoring of NPS. Additionally, UHPSFC has the potential to combine the advantages of LC  
656 and GC, thus improving analytical capabilities of laboratories dealing with the determination of  
657 NPS.

658 iv. Highly sensitive **targeted methodologies** based on LRMS will continue to play an important role  
659 in monitoring NPS use, particularly for those compounds which have established a niche market  
660 and/or are highly potent and require low detection limits. In addition, complementary **suspect**  
661 **screening approaches** based on large home-made databases, including many substances for  
662 which reference standards are not available, will remain the common practice for the foreseeable



663 future. Furthermore, the improved sensitivity and quantitative capabilities of HRMS instruments  
664 combined to multi-stage off-line or on-line solid-phase extraction allow achieving targeted  
665 quantitative and qualitative screening analyses in a single run, thus overcoming the need of having  
666 two distinct instruments/methods [27]. Similarly, machine learning algorithms used to relate peak  
667 area of features recorded in HRMS analyses, chromatographic and mass spectrometric conditions  
668 to concentrations, might overcome the need for reference standards to obtain an (indicative)  
669 information about analyte concentrations in measured samples [135]. Qualitative information  
670 about the presence or absence of given NPS in wastewater is informative and studies have shown  
671 some spatial and temporal trends [23,27,136], but only quantitative data can provide absolute  
672 comparisons by showing changes in community prevalence through concentrations or mass loads.

673 v. **Non-target screening** remains predominantly unexplored for the identification of NPS in pooled  
674 urine or wastewater. A genuine non-target screening without any selection of analytes to be  
675 searched is a very challenging and time consuming process and a more successful strategy would  
676 be the application of non-target screening directed towards the discovery of compounds  
677 structurally related to known NPS. In this case, the higher concentrations generally present in  
678 pooled urine makes this matrix most interesting for this approach. The expected improvements  
679 for the forthcoming years in the mass-resolving power of HRMS instruments in combination with  
680 higher scan-speed will allow the acquisition at higher mass resolution with more efficient  
681 chromatography. This development in instrumentation will improve sensitivity and can also be  
682 very useful to differentiate between isobaric compounds (i.e. compounds with the same nominal  
683 mass but different chemical formula and thus different exact mass). Moreover, improved mass  
684 resolving power does not only improve the separation of parent compounds, but can also help  
685 finding characteristic fragment ions and gain confidence in the obtain identification. Furthermore,  
686 improvements in software tools for peak picking and data deconvolution (*i.e.* the capability to find  
687 chromatographic peaks of compounds and to obtain high quality spectra) will aid to a successful  
688 identification of NPS, but the knowledge of basic rules in mass fragmentation and thus the  
689 expertise of the mass spectrometrists should not be overlooked in both suspect and non-target  
690 screening.

691 vi. The rapid transience of NPS in the drug market as well as the limited availability of reference  
692 standards for both NPS and known metabolites poses an analytical challenge for the full  
693 confirmation of substances detected. Therefore, the development and continuous updating of  
694 **collaborative and public NPS mass spectral databases** will smooth the identification process since  
695 contributors and users to those databases will have access to empirical information without the  
696 need of having the reference standards in their own laboratories. Hence, the number of false

697 positive identification (based on suspect and non-target screening) will be reduced since tentative  
698 identifications will be supported by empirical data from other researchers.

699 vii. As is the case with online databases, **prediction tools** ease the tentative identification of NPS. The  
700 development of metabolic, RT and CCS predictive models represent a turning point in the  
701 investigation of NPS. The continual development of more accurate and refined predictive models  
702 will make prediction tools even more powerful for the application of NPS consumption –  
703 particularly the complexity associated with structural similarities among NPS families. The small  
704 differences in the chemical backbone for most NPS classes and consequently similar  
705 physicochemical properties often make the current predictive tools less than ideal due to the  
706 analogous outcome obtained from the prediction.

707 viii. **Retrospective analysis** will continue to play an important role in uncovering trends in NPS  
708 consumption. HRMS analyses allow analysts to continually explore samples, without the time  
709 expense associate with re-extracting and re-analyzing samples. Reprocessing samples should be  
710 performed periodically, which can be a laborious task. Nevertheless, it is an interesting tool, as  
711 ‘new’ NPS and metabolites are found, standards become more available and predictive techniques  
712 become more commonplace, retrospective analyses can be performed to better reveal  
713 community use of NPS.

714 ix. **Ion mobility separation** coupled to HRMS has arisen as a useful technique and it is expected that  
715 it will gain in popularity. The cleaner and higher-quality mass spectra as well as the increased  
716 sensitivity of the instruments facilitates the identification process of NPS at low concentration  
717 levels and in complex wastewater or pooled urine samples. Future improvements will be related  
718 to the resolution of IMS instrument to enhance the separation of isobaric or isomeric substances  
719 that cannot be previously resolved by chromatography.

## 720 **8. Conclusions**

721 Comprehensive analytical strategies can be applied to investigate NPS in pooled urine and  
722 wastewater, from quantification of target biomarkers to the detection and (tentative) identification  
723 of new substances and metabolites. The investigation of NPS in pooled urine and wastewater is a  
724 subject of current interest because, integrated with additional epidemiological information, it can be  
725 a useful tool for a comprehensive assessment of NPS use. In this context, data triangulation with  
726 traditional indicators, such as public surveys, online forums, data of drug testing services, police  
727 seizures and forensic analyses, is pivotal to gauge community consumption. Thus, the analysis of  
728 pooled urine and wastewater can complement other data and provide a more complete picture of  
729 community consumption.

730

## 731 **Acknowledgements**

732 The authors acknowledge EuSeME (project number 861602) funded by the European Union's Justice  
733 Programme - Drugs Policy Initiatives, and NPS-Euronet (HOME/2014/JDRUG/AG/DRUG/7086) funded  
734 with support from the European Commission. This communication reflects the views only of the  
735 authors, and the European Commission cannot be held responsible for any use that may be made of  
736 the information contained therein. Alberto Celma acknowledges the Spanish Ministry for Economy  
737 and Competitiveness for his pre-doctoral grant (BES-2016-076914). Richard Bade acknowledges the  
738 financial support of the Thyne Reid Foundation. The authors also acknowledge Juan Vicente Sancho  
739 of the Research Institute for Pesticides and Water for his help and fruitful discussions, Florenci Vicent  
740 González Adelantado and Lledo Bou-Iserte of the Department of Inorganic and Organic Chemistry,  
741 University Jaume I (Spain) for the synthesis of  $\alpha$ -methyltryptamine.

742 **References**

743

- 744 [1] European Monitoring Centre for Drugs and Drug Addiction, European Drug Report 2019:  
745 Trends and Developments, Lisbon, 2019.  
746 [http://www.emcdda.europa.eu/system/files/publications/4541/TDAT17001ENN.pdf\\_en](http://www.emcdda.europa.eu/system/files/publications/4541/TDAT17001ENN.pdf_en).
- 747 [2] United Nations Office on Drugs and Crime, Analysis of drug markets. World Drug Report  
748 2018, 2018. doi:10.18356/dc023cb1-en.
- 749 [3] R. Bade, P. Stockham, B. Painter, A. Celma, L. Bijlsma, F. Hernandez, J.M. White, C. Gerber,  
750 Investigating the appearance of new psychoactive substances in South Australia using  
751 wastewater and forensic data, *Drug Test. Anal.* 11 (2019) 250–256. doi:10.1002/dta.2484.
- 752 [4] B. Hughes, J. Matias, P. Griffiths, Inconsistencies in the assumptions linking punitive sanctions  
753 and use of cannabis and new psychoactive substances in Europe, *Addiction*. (2018) 2155–  
754 2157. doi:10.1111/add.14372.
- 755 [5] J.B. Zawilska, “Legal Highs” - An emerging epidemic of novel psychoactive substances, 1st ed.,  
756 Elsevier Inc., 2015. doi:10.1016/bs.irn.2015.02.009.
- 757 [6] A.J. Krotulski, D.M. Papsun, M. Friscia, J.L. Swartz, B.D. Holsey, B.K. Logan, Fatality following  
758 ingestion of tetrahydrofuranlyfentanyl, U-49900 and methoxy-phencyclidine, *J. Anal. Toxicol.*  
759 42 (2018) e27–e32. doi:10.1093/jat/bkx092.
- 760 [7] European Monitoring Centre for Drugs and Drug Addiction, EMCDDA: The EU early warning  
761 system, *Eur. Monit. Cent. Drugs Drugs Addict.* (2018).  
762 <http://www.emcdda.europa.eu/themes/new-drugs/early-warning> (accessed October 15,  
763 2019).
- 764 [8] R.A.S. Couto, L.M. Gonçalves, F. Carvalho, J.A. Rodrigues, C.M.P. Rodrigues, M.B. Quinaz, The  
765 Analytical Challenge in the Determination of Cathinones, Key-Players in the Worldwide  
766 Phenomenon of Novel Psychoactive Substances, *Crit. Rev. Anal. Chem.* 48 (2018) 372–390.  
767 doi:10.1080/10408347.2018.1439724.
- 768 [9] V. Angerer, L. Mogler, J.P. Steitz, P. Bisel, C. Hess, C.T. Schoeder, C.E. Müller, L.M. Huppertz, F.  
769 Westphal, J. Schäper, V. Auwärter, Structural characterization and pharmacological  
770 evaluation of the new synthetic cannabinoid CUMYL-PEGACLONE, *Drug Test. Anal.* 10 (2018)  
771 597–603. doi:10.1002/dta.2237.
- 772 [10] N. Uchiyama, Y. Shimokawa, M. Kawamura, R. Kikura-Hanajiri, T. Hakamatsuka, Chemical  
773 analysis of a benzofuran derivative, 2-(2-ethylaminopropyl) benzofuran (2-EAPB), eight  
774 synthetic cannabinoids, five cathinone derivatives, and five other designer drugs newly  
775 detected in illegal products, *Forensic Toxicol.* 32 (2014) 266–281. doi:10.1007/s11419-014-  
776 0238-5.
- 777 [11] L. Bijlsma, B. Miserez, M. Ibáñez, C. Vicent, E. Guillamón, J. Ramsey, F. Hernández,  
778 Identification and characterization of a novel cathinone derivative 1-(2,3-dihydro-1H-inden-5-  
779 yl)-2-phenyl-2-(pyrrolidin-1-yl)-ethanone seized by customs in Jersey, *Forensic Toxicol.* 34  
780 (2016) 144–150. doi:10.1007/s11419-015-0299-0.
- 781 [12] E. Partridge, S. Trobbiani, P. Stockham, T. Scott, C. Kostakis, A Validated Method for the  
782 Screening of 320 Forensically Significant Compounds in Blood by LC/QTOF, with Simultaneous  
783 Quantification of Selected Compounds, *J. Anal. Toxicol.* 42 (2018) 220–231.  
784 doi:10.1093/jat/bkx108.

- 785 [13] M.R. Meyer, H.H. Maurer, LC coupled to low- and high-resolution mass spectrometry for new  
786 psychoactive substance screening in biological matrices – Where do we stand today?, *Anal.*  
787 *Chim. Acta.* 927 (2016) 13–20. doi:10.1016/j.aca.2016.04.046.
- 788 [14] D. Pasin, A. Cawley, S. Bidny, S. Fu, Current applications of high-resolution mass spectrometry  
789 for the analysis of new psychoactive substances: a critical review, *Anal. Bioanal. Chem.* 409  
790 (2017) 5821–5836. doi:10.1007/s00216-017-0441-4.
- 791 [15] I. González-Mariño, E. Gracia-Lor, N.I. Rousis, E. Castrignanò, K. V. Thomas, J.B. Quintana, B.  
792 Kasprzyk-Hordern, E. Zuccato, S. Castiglioni, Wastewater-Based Epidemiology To Monitor  
793 Synthetic Cathinones Use in Different European Countries, *Environ. Sci. Technol.* 50 (2016)  
794 10089–10096. doi:10.1021/acs.est.6b02644.
- 795 [16] C. Ort, L. Bijlsma, S. Castiglioni, A. Covaci, P. de Voogt, E. Emke, F. Hernández, M. Reid, A.L.N.  
796 van Nuijs, K. V. Thomas, B. Kasprzyk-Hordern, Wastewater Analysis for Community-Wide  
797 Drugs Use Assessment, in: H.H. Maurer, S.D. Brandt (Eds.), *Handb. Exp. Pharmacol.*, Springer  
798 Berlin Heidelberg, 2018: pp. 1–24. doi:10.1007/164\_2018\_111.
- 799 [17] L. Bijlsma, A. Celma, F.J. López, F. Hernández, Monitoring new psychoactive substances use  
800 through wastewater analysis: current situation, challenges and limitations, *Curr. Opin.*  
801 *Environ. Sci. Heal.* 9 (2019) 1–12. doi:10.1016/j.coesh.2019.03.002.
- 802 [18] J.R.H. Archer, P.I. Dargan, S. Hudson, D.M. Wood, Analysis of anonymous pooled urine from  
803 portable urinals in central london confirms the significant use of novel psychoactive  
804 substances, *QJM An Int. J. Med.* 106 (2013) 147–152. doi:10.1093/qjmed/hcs219.
- 805 [19] J.R.H. Archer, S. Hudson, O. Jackson, T. Yamamoto, C. Lovett, H.M. Lee, S. Rao, L. Hunter, P.I.  
806 Dargan, D.M. Wood, Analysis of anonymized pooled urine in nine UK cities: Variation in  
807 classical recreational drug, novel psychoactive substance and anabolic steroid use, *QJM An*  
808 *Int. J. Med.* 108 (2015) 929–933. doi:10.1093/qjmed/hcv058.
- 809 [20] J. Kinyua, N. Negreira, B. Miserez, A. Causanilles, E. Emke, L. Gremeaux, P. de Voogt, J.  
810 Ramsey, A. Covaci, A.L.N. van Nuijs, Qualitative screening of new psychoactive substances in  
811 pooled urine samples from Belgium and United Kingdom, *Sci. Total Environ.* 573 (2016) 1527–  
812 1535. doi:10.1016/j.scitotenv.2016.08.124.
- 813 [21] J.R.H. Archer, P.I. Dargan, H.M.D. Lee, S. Hudson, D.M. Wood, Trend analysis of anonymised  
814 pooled urine from portable street urinals in central London identifies variation in the use of  
815 novel psychoactive substances, *Clin. Toxicol.* 52 (2014) 160–165.  
816 doi:10.3109/15563650.2014.885982.
- 817 [22] H. Gjerde, L. Gjersing, J.A. Baz-Lomba, L. Bijlsma, N. Salgueiro-González, H. Furuhaugen, A.L.  
818 Bretteville-Jensen, F. Hernández, S. Castiglioni, E. Johanna Amundsen, E. Zuccato, Drug Use by  
819 Music Festival Attendees: A Novel Triangulation Approach Using Self-Reported Data and Test  
820 Results of Oral Fluid and Pooled Urine Samples, *Subst. Use Misuse.* 0 (2019) 1–11.  
821 doi:10.1080/10826084.2019.1646285.
- 822 [23] L. Bijlsma, A. Celma, S. Castiglioni, N. Salgueiro-González, L. Bou-Iserte, J.A.J.A. Baz-Lomba,  
823 M.J.M.J. Reid, M.J.M.J. Dias, A. Lopes, J. Matias, L. Pastor-Alcañiz, J. Radonic, M.T. Sekulic, T.  
824 Shine, A.L.N.A.L.N. van Nuijs, F. Hernández, E. Zuccato, J. Radonić, M. Turk Sekulic, T. Shine,  
825 A.L.N.A.L.N. van Nuijs, F. Hernandez, E. Zuccato, Monitoring psychoactive substances use at  
826 six European festivals through wastewater and pooled urine analysis, *Sci. Total Environ.* 725  
827 (2020). doi:10.1016/j.scitotenv.2020.138376.
- 828 [24] L. Benaglia, R. Udrisard, A. Bannwarth, A. Gibson, F. Béen, F.Y. Lai, P. Esseiva, O. Delémont,  
829 Testing wastewater from a music festival in Switzerland to assess illicit drug use, *Forensic Sci.*

- 830 Int. 309 (2020) 1–8. doi:10.1016/j.forsciint.2020.110148.
- 831 [25] J.R.H. Archer, P.I. Dargan, S. Hudson, S. Davies, M. Puchnarewicz, A.T. Kicman, J. Ramsey, F.  
832 Measham, M. Wood, A. Johnston, D.M. Wood, Taking the Piss - A novel and reliable way of  
833 knowing what drugs are being used in nightclubs, *J. Subst. Use.* 19 (2014) 103–107.  
834 doi:10.3109/14659891.2012.740139.
- 835 [26] A. Celma, J. V. Sancho, N. Salgueiro-González, S. Castiglioni, E. Zuccato, F. Hernández, L.  
836 Bijlsma, Simultaneous determination of new psychoactive substances and illicit drugs in  
837 sewage: Potential of micro-liquid chromatography tandem mass spectrometry in wastewater-  
838 based epidemiology, *J. Chromatogr. A.* 1602 (2019) 300–309.  
839 doi:10.1016/j.chroma.2019.05.051.
- 840 [27] K. Diamanti, R. Aalizadeh, N. Alygizakis, A. Galani, M. Mardal, N.S. Thomaidis, Wide-scope  
841 target and suspect screening methodologies to investigate the occurrence of new  
842 psychoactive substances in influent wastewater from Athens, *Sci. Total Environ.* 685 (2019)  
843 1058–1065. doi:10.1016/j.scitotenv.2019.06.173.
- 844 [28] C. Ort, A.L.N. van Nuijs, J.D. Berset, L. Bijlsma, S. Castiglioni, A. Covaci, P. de Voogt, E. Emke,  
845 D. Fatta-Kassinos, P. Griffiths, F. Hernández, I. González-Mariño, R. Grabic, B. Kasprzyk-  
846 Hordern, N. Mastroianni, A. Meierjohann, T. Nefau, M. Östman, Y. Pico, I. Racamonde, M.  
847 Reid, J. Slobodnik, S. Terzic, N. Thomaidis, K. V. Thomas, Spatial differences and temporal  
848 changes in illicit drug use in Europe quantified by wastewater analysis, *Addiction.* 109 (2014)  
849 1338–1352. doi:10.1111/add.12570.
- 850 [29] I. González-Mariño, J.A. Baz-Lomba, N.A. Alygizakis, M.J. Andrés-Costa, R. Bade, A.  
851 Bannwarth, L.P. Barron, F. Been, L. Benaglia, J. Berset, L. Bijlsma, I. Bodík, A. Brenner, A.L.  
852 Brock, D.A. Burgard, E. Castrignanò, A. Celma, C.E. Christophoridis, A. Covaci, O. Delémont, P.  
853 Voogt, D.A. Devault, M.J. Dias, E. Emke, P. Esseiva, D. Fatta-Kassinos, G. Fedorova, K. Fytianos,  
854 C. Gerber, R. Grabic, E. Gracia-Lor, S. Grüner, T. Gunnar, E. Hapeshi, E. Heath, B. Helm, F.  
855 Hernández, A. Kankaanpaa, S. Karolak, B. Kasprzyk-Hordern, I. Krizman-Matasic, F.Y. Lai, W.  
856 Lechowicz, A. Lopes, M. López de Alda, E. López-García, A.S.C.C. Löve, N. Mastroianni, G.L.  
857 McEneff, R. Montes, K. Munro, T. Nefau, H. Oberacher, J.W. O'Brien, R. Oertel, K. Olafsdottir,  
858 Y. Picó, B.G. Plósz, F. Polesel, C. Postigo, J.B. Quintana, P. Ramin, M.J. Reid, J. Rice, R. Rodil, N.  
859 Salgueiro-González, S. Schubert, I. Senta, S.M. Simões, M.M. Sremacki, K. Styszko, S. Terzic,  
860 N.S. Thomaidis, K. V. Thomas, B.J. Tschärke, R. Udrisard, A.L.N. Nuijs, V. Yargeau, E. Zuccato,  
861 S. Castiglioni, C. Ort, Spatio-temporal assessment of illicit drug use at large scale: evidence  
862 from 7 years of international wastewater monitoring, *Addiction.* 115 (2020) 109–120.  
863 doi:10.1111/add.14767.
- 864 [30] E. Zuccato, C. Chiabrando, S. Castiglioni, R. Bagnati, R. Fanelli, Estimating community drug  
865 abuse by wastewater analysis, *Environ. Health Perspect.* 116 (2008) 1027–1032.  
866 doi:10.1289/ehp.11022.
- 867 [31] S. Castiglioni, L. Bijlsma, A. Covaci, E. Emke, F. Hernández, M. Reid, C. Ort, K. V. Thomas,  
868 A.L.N. Van Nuijs, P. De Voogt, E. Zuccato, Evaluation of uncertainties associated with the  
869 determination of community drug use through the measurement of sewage drug biomarkers,  
870 *Environ. Sci. Technol.* 47 (2013) 1452–1460. doi:10.1021/es302722f.
- 871 [32] European Monitoring Centre for Drugs and Drug Addiction, Assessing illicit drugs in  
872 wastewater, EMCDDA Ins, Publications Office of the European Union, Luxembourg, 2016.  
873 doi:10.2810/017397.
- 874 [33] J.A. Baz-Lomba, C. Harman, M. Reid, K. V. Thomas, Passive sampling of wastewater as a tool  
875 for the long-term monitoring of community exposure: Illicit and prescription drug trends as a

- 876 proof of concept, *Water Res.* 121 (2017) 221–230. doi:10.1016/j.watres.2017.05.041.
- 877 [34] C. Harman, I.J. Allan, E.L.M. Vermeirssen, Calibration and use of the polar organic chemical  
878 integrative sampler—a critical review, *Environ. Toxicol. Chem.* 31 (2012) 2724–2738.  
879 doi:10.1002/etc.2011.
- 880 [35] P.O.M. Gundersen, O. Spigset, M. Josefsson, Screening, quantification, and confirmation of  
881 synthetic cannabinoid metabolites in urine by UHPLC-QTOF-MS, *Drug Test. Anal.* 11 (2019)  
882 51–67. doi:10.1002/dta.2464.
- 883 [36] K.A. Alsenedi, C. Morrison, Determination and long-term stability of twenty-nine cathinones  
884 and amphetamine-type stimulants (ATS) in urine using gas chromatography–mass  
885 spectrometry, *J. Chromatogr. B.* 1076 (2018) 91–102. doi:10.1016/j.jchromb.2018.01.027.
- 886 [37] P. Adamowicz, A. Malczyk, Stability of synthetic cathinones in blood and urine, *Forensic Sci.*  
887 *Int.* 295 (2019) 36–45. doi:10.1016/j.forsciint.2018.12.001.
- 888 [38] M. Pettersson Bergstrand, O. Beck, A. Helander, Urine analysis of 28 designer  
889 benzodiazepines by liquid chromatography–high-resolution mass spectrometry, *Clin. Mass*  
890 *Spectrom.* 10 (2018) 25–32. doi:10.1016/j.clinms.2018.08.004.
- 891 [39] R. Bade, A. Abdelaziz, L. Nguyen, A.J. Pandopulos, J.M. White, C. Gerber, Determination of 21  
892 synthetic cathinones, phenethylamines, amphetamines and opioids in influent wastewater  
893 using liquid chromatography coupled to tandem mass spectrometry, *Talanta.* 208 (2020).  
894 doi:10.1016/j.talanta.2019.120479.
- 895 [40] R. Bade, L. Bijlsma, J. V. Sancho, J.A. Baz-Lomba, S. Castiglioni, E. Castrignanò, A. Causanilles,  
896 E. Gracia-Lor, B. Kasprzyk-Hordern, J. Kinyua, A.K. McCall, A.L.N. van Nuijs, C. Ort, B.G. Plósz,  
897 P. Ramin, N.I. Rousis, Y. Ryu, K. V. Thomas, P. de Voogt, E. Zuccato, F. Hernández, Liquid  
898 chromatography-tandem mass spectrometry determination of synthetic cathinones and  
899 phenethylamines in influent wastewater of eight European cities, *Chemosphere.* 168 (2017)  
900 1032–1041. doi:10.1016/j.chemosphere.2016.10.107.
- 901 [41] I. Senta, I. Krizman, M. Ahel, S. Terzic, Multiresidual analysis of emerging amphetamine-like  
902 psychoactive substances in wastewater and river water, *J. Chromatogr. A.* 1425 (2015) 204–  
903 212. doi:10.1016/j.chroma.2015.11.043.
- 904 [42] M.J. Reid, L. Derry, K. V. Thomas, Analysis of new classes of recreational drugs in sewage:  
905 Synthetic cannabinoids and amphetamine-like substances, *Drug Test. Anal.* 6 (2014) 72–79.  
906 doi:10.1002/dta.1461.
- 907 [43] A.J. Pandopulos, R. Bade, J.W. O’Brien, B.J. Tschärke, J.F. Mueller, K. Thomas, J.M. White, C.  
908 Gerber, Towards an efficient method for the extraction and analysis of cannabinoids in  
909 wastewater, *Talanta.* 217 (2020) 121034. doi:10.1016/j.talanta.2020.121034.
- 910 [44] L.C.G. Hoegberg, C. Christiansen, J. Soe, R. Telving, M.F. Andreasen, D. Staerk, L.L. Christrup,  
911 K.T. Kongstad, Recreational drug use at a major music festival: trend analysis of anonymised  
912 pooled urine, *Clin. Toxicol.* 56 (2018) 245–255. doi:10.1080/15563650.2017.1360496.
- 913 [45] M. Mardal, J. Kinyua, P. Ramin, B. Miserez, A.L.N. Van Nuijs, A. Covaci, M.R. Meyer, Screening  
914 for illicit drugs in pooled human urine and urinated soil samples and studies on the stability of  
915 urinary excretion products of cocaine, MDMA, and MDEA in wastewater by hyphenated mass  
916 spectrometry techniques, *Drug Test. Anal.* 9 (2017) 106–114. doi:10.1002/dta.1957.
- 917 [46] V.L. Borova, P. Gago-Ferrero, C. Pistos, N.S. Thomaidis, Multi-residue determination of 10  
918 selected new psychoactive substances in wastewater samples by liquid chromatography-  
919 tandem mass spectrometry, *Talanta.* 144 (2015) 592–603. doi:10.1016/j.talanta.2015.06.080.

- 920 [47] M.R. Boleda, M.T. Galceran, F. Ventura, Trace determination of cannabinoids and opiates in  
921 wastewater and surface waters by ultra-performance liquid chromatography–tandem mass  
922 spectrometry, *J. Chromatogr. A.* 1175 (2007) 38–48. doi:10.1016/j.chroma.2007.10.029.
- 923 [48] S. Castiglioni, E. Zuccato, E. Crisci, C. Chiabrando, R. Fanelli, R. Bagnati, Identification and  
924 measurement of illicit drugs and their metabolites in urban wastewater by liquid  
925 chromatography-tandem mass spectrometry, *Anal. Chem.* 78 (2006) 8421–8429.  
926 doi:10.1021/ac061095b.
- 927 [49] J. Gao, A. Banks, J. Li, G. Jiang, F.Y. Lai, J.F. Mueller, P.K. Thai, Evaluation of in-sewer  
928 transformation of selected illicit drugs and pharmaceutical biomarkers, *Sci. Total Environ.* 609  
929 (2017) 1172–1181. doi:10.1016/j.scitotenv.2017.07.231.
- 930 [50] I. Krizman-Matasic, P. Kostanjevecki, M. Ahel, S. Terzic, Simultaneous analysis of opioid  
931 analgesics and their metabolites in municipal wastewaters and river water by liquid  
932 chromatography–tandem mass spectrometry, *J. Chromatogr. A.* 1533 (2018) 102–111.  
933 doi:10.1016/j.chroma.2017.12.025.
- 934 [51] M. Mardal, M.R. Meyer, Studies on the microbial biotransformation of the novel psychoactive  
935 substance methylenedioxypropylvalerone (MDPV) in wastewater by means of liquid  
936 chromatography-high resolution mass spectrometry/mass spectrometry, *Sci. Total Environ.*  
937 493 (2014) 588–595. doi:10.1016/j.scitotenv.2014.06.016.
- 938 [52] N. Salgueiro-González, S. Castiglioni, E. Gracia-Lor, L. Bijlsma, A. Celma, R. Bagnati, F.  
939 Hernández, E. Zuccato, Flexible high resolution-mass spectrometry approach for screening  
940 new psychoactive substances in urban wastewater, *Sci. Total Environ.* 689 (2019) 679–690.  
941 doi:10.1016/j.scitotenv.2019.06.336.
- 942 [53] M.J. Andrés-Costa, V. Andreu, Y. Picó, Liquid chromatography–mass spectrometry as a tool  
943 for wastewater-based epidemiology: Assessing new psychoactive substances and other  
944 human biomarkers, *TrAC - Trends Anal. Chem.* 94 (2017) 21–38.  
945 doi:10.1016/j.trac.2017.06.012.
- 946 [54] C.E. O'Rourke, B. Subedi, Occurrence and Mass Loading of Synthetic Opioids, Synthetic  
947 Cathinones, and Synthetic Cannabinoids in Wastewater Treatment Plants in Four U.S.  
948 Communities, *Environ. Sci. Technol.* 54 (2020) 6661–6670. doi:10.1021/acs.est.0c00250.
- 949 [55] J. Kinyua, A. Covaci, W. Maho, A.K. Mccall, H. Neels, A.L.N. van Nuijs, Sewage-based  
950 epidemiology in monitoring the use of new psychoactive substances: Validation and  
951 application of an analytical method using LC-MS/MS, *Drug Test. Anal.* 7 (2015) 812–818.  
952 doi:10.1002/dta.1777.
- 953 [56] E. López-García, N. Mastroianni, C. Postigo, D. Barceló, M. López de Alda, A fully automated  
954 approach for the analysis of 37 psychoactive substances in raw wastewater based on on-line  
955 solid phase extraction-liquid chromatography-tandem mass spectrometry, *J. Chromatogr. A.*  
956 1576 (2018) 80–89. doi:10.1016/j.chroma.2018.09.038.
- 957 [57] J.R.H. Archer, F. Mendes, S. Hudson, K. Layne, P.I. Dargan, D.M. Wood, Evaluation of long-  
958 term detection trends of new psychoactive substances in pooled urine from city street  
959 portable urinals (London, UK), *Br. J. Clin. Pharmacol.* 86 (2020) 517–527.  
960 doi:10.1111/bcp.14239.
- 961 [58] I. González-Mariño, K. V. Thomas, M.J. Reid, Determination of cannabinoid and synthetic  
962 cannabinoid metabolites in wastewater by liquid–liquid extraction and ultra-high  
963 performance supercritical fluid chromatography-tandem mass spectrometry, *Drug Test. Anal.*  
964 10 (2018) 222–228. doi:10.1002/dta.2199.



- 965 [59] I. Marchi, V. Viette, F. Badoud, M. Fathi, M. Saugy, S. Rudaz, J.L. Veuthey, Characterization  
966 and classification of matrix effects in biological samples analyses, *J. Chromatogr. A.* 1217  
967 (2010) 4071–4078. doi:10.1016/j.chroma.2009.08.061.
- 968 [60] L. Bijlsma, E. Beltrán, C. Boix, J.V. Sancho, F. Hernández, Improvements in analytical  
969 methodology for the determination of frequently consumed illicit drugs in urban wastewater,  
970 *Anal. Bioanal. Chem.* 406 (2014) 4261–4272. doi:10.1007/s00216-014-7818-4.
- 971 [61] J.Y. Kim, S. Suh, J. Park, M.K. In, Simultaneous Determination of Amphetamine-Related New  
972 Psychoactive Substances in Urine by Gas Chromatography–Mass Spectrometry†, *J. Anal.*  
973 *Toxicol.* 42 (2018) 605–616. doi:10.1093/jat/bky037.
- 974 [62] L.A. Nisbet, F.M. Wylie, B.K. Logan, K.S. Scott, Gas Chromatography-Mass Spectrometry  
975 Method for the Quantitative Identification of 23 New Psychoactive Substances in Blood and  
976 Urine, *J. Anal. Toxicol.* 43 (2019) 346–352. doi:10.1093/jat/bky109.
- 977 [63] F. Hernández, S. Castiglioni, A. Covaci, P. de Voogt, E. Emke, B. Kasprzyk-Hordern, C. Ort, M.  
978 Reid, J. V. Sancho, K. V. Thomas, A.L.N. van Nuijs, E. Zuccato, L. Bijlsma, Mass spectrometric  
979 strategies for the investigation of biomarkers of illicit drug use in wastewater, *Mass*  
980 *Spectrom. Rev.* 37 (2018) 258–280. doi:10.1002/mas.21525.
- 981 [64] M. Concheiro, M. Castaneto, R. Kronstrand, M.A. Huestis, Simultaneous determination of 40  
982 novel psychoactive stimulants in urine by liquid chromatography-high resolution mass  
983 spectrometry and library matching, *J. Chromatogr. A.* 1397 (2015) 32–42.  
984 doi:10.1016/j.chroma.2015.04.002.
- 985 [65] I. González-Mariño, V. Castro, R. Montes, R. Rodil, A. Lores, R. Cela, J.B. Quintana, Multi-  
986 residue determination of psychoactive pharmaceuticals, illicit drugs and related metabolites  
987 in wastewater by ultra-high performance liquid chromatography-tandem mass spectrometry,  
988 *J. Chromatogr. A.* 1569 (2018) 91–100. doi:10.1016/j.chroma.2018.07.045.
- 989 [66] E. Castrignanò, M. Mardal, A. Rydevik, B. Miserez, J. Ramsey, T. Shine, G.D. Panto, M.R.  
990 Meyer, B. Kasprzyk-Hordern, A new approach towards biomarker selection in estimation of  
991 human exposure to chiral chemicals: A case study of mephedrone, *Sci. Rep.* 7 (2017) 1–12.  
992 doi:10.1038/s41598-017-12581-3.
- 993 [67] E. Castrignanò, A. Lubben, B. Kasprzyk-Hordern, Enantiomeric profiling of chiral drug  
994 biomarkers in wastewater with the usage of chiral liquid chromatography coupled with  
995 tandem mass spectrometry, *J. Chromatogr. A.* 1438 (2016) 84–99.  
996 doi:10.1016/j.chroma.2016.02.015.
- 997 [68] E. Castrignanò, Z. Yang, R. Bade, J.A. Baz-Lomba, S. Castiglioni, A. Causanilles, A. Covaci, E.  
998 Gracia-Lor, F. Hernandez, J. Kinyua, A.-K. McCall, A.L.N. van Nuijs, C. Ort, B.G. Plósz, P. Ramin,  
999 N.I. Rousis, Y. Ryu, K. V. Thomas, P. de Voogt, E. Zuccato, B. Kasprzyk-Hordern, Enantiomeric  
1000 profiling of chiral illicit drugs in a pan-European study, *Water Res.* 130 (2018) 151–160.  
1001 doi:10.1016/j.watres.2017.11.051.
- 1002 [69] B. Kasprzyk-Hordern, D.R. Baker, Estimation of community-wide drugs use via stereoselective  
1003 profiling of sewage, *Sci. Total Environ.* 423 (2012) 142–150.  
1004 doi:10.1016/j.scitotenv.2012.02.019.
- 1005 [70] K.H. Storbeck, L. Gilligan, C. Jenkinson, E.S. Baranowski, J.L. Quanson, W. Arlt, A.E. Taylor, The  
1006 utility of ultra-high performance supercritical fluid chromatography–tandem mass  
1007 spectrometry (UHPSFC-MS/MS) for clinically relevant steroid analysis, *J. Chromatogr. B Anal.*  
1008 *Technol. Biomed. Life Sci.* 1085 (2018) 36–41. doi:10.1016/j.jchromb.2018.03.033.

- 1009 [71] L. Nováková, A. Grand-Guillaume Perrenoud, I. Francois, C. West, E. Lesellier, D. Guillarme,  
1010 Modern analytical supercritical fluid chromatography using columns packed with sub-2 $\mu$ m  
1011 particles: A tutorial, *Anal. Chim. Acta.* 824 (2014) 18–35. doi:10.1016/j.aca.2014.03.034.
- 1012 [72] Å.M. Leere Øiestad, T. Berg, E. Eliassen, T. Wiklund, K. Sand, E. Leere Øiestad, Å.M.L. Øiestad,  
1013 T. Berg, E. Eliassen, T. Wiklund, K. Sand, E. Leere Øiestad, Separation of isomers of new  
1014 psychoactive substances and isotope-labeled amphetamines using UHPSFC-MS/MS and  
1015 UHPLC-MS/MS, *J. Liq. Chromatogr. Relat. Technol.* 41 (2018) 391–400.  
1016 doi:10.1080/10826076.2017.1388818.
- 1017 [73] A.L.N. van Nuijs, A. Gheorghe, P.G. Jorens, K. Maudens, H. Neels, A. Covaci, Optimization,  
1018 validation, and the application of liquid chromatography-tandem mass spectrometry for the  
1019 analysis of new drugs of abuse in wastewater, *Drug Test. Anal.* 6 (2014) 861–867.  
1020 doi:10.1002/dta.1460.
- 1021 [74] A. Kankaanpää, K. Ariniemi, M. Heinonen, K. Kuoppasalmi, T. Gunnar, Use of illicit stimulant  
1022 drugs in Finland: A wastewater study in ten major cities, *Sci. Total Environ.* 487 (2014) 696–  
1023 702. doi:10.1016/j.scitotenv.2013.11.095.
- 1024 [75] European Commission, Commission Decision of 12 August 2002 implementing Council  
1025 Directive 96/23/EC concerning the performance of analytical methods and the interpretation  
1026 of results (2002/657/EC), 2002.
- 1027 [76] European Commission. Directorate General for Health and Food Safety., Guidance document  
1028 on analytical quality control and method validation procedures for pesticide residue and  
1029 analysis in food and feed. SANTE/11813/2017, 2017.
- 1030 [77] M. Krauss, H. Singer, J. Hollender, LC-high resolution MS in environmental analysis: From  
1031 target screening to the identification of unknowns, *Anal. Bioanal. Chem.* 397 (2010) 943–951.  
1032 doi:10.1007/s00216-010-3608-9.
- 1033 [78] T. Gao, P. Du, Z. Xu, X. Li, Occurrence of new psychoactive substances in wastewater of major  
1034 Chinese cities, *Sci. Total Environ.* 575 (2017) 963–969. doi:10.1016/j.scitotenv.2016.09.152.
- 1035 [79] B.J. Tschärke, C. Chen, J.P. Gerber, J.M. White, Temporal trends in drug use in Adelaide, South  
1036 Australia by wastewater analysis, *Sci. Total Environ.* 565 (2016) 384–391.  
1037 doi:10.1016/j.scitotenv.2016.04.183.
- 1038 [80] M. Mardal, B. Miserez, R. Bade, T. Portolés, M. Bischoff, F. Hernández, M.R. Meyer, 3-  
1039 Fluorophenmetrazine, a fluorinated analogue of phenmetrazine: Studies on in vivo  
1040 metabolism in rat and human, in vitro metabolism in human CYP isoenzymes and microbial  
1041 biotransformation in *Pseudomonas Putida* and wastewater using GC and LC coupled to (HR), *J.*  
1042 *Pharm. Biomed. Anal.* 128 (2016) 485–495. doi:10.1016/j.jpba.2016.06.011.
- 1043 [81] F.Y. Lai, C. Erratico, J. Kinyua, J.F. Mueller, A. Covaci, A.L.N. van Nuijs, Liquid chromatography-  
1044 quadrupole time-of-flight mass spectrometry for screening in vitro drug metabolites in  
1045 humans: investigation on seven phenethylamine-based designer drugs, *J. Pharm. Biomed.*  
1046 *Anal.* 114 (2015) 355–375. doi:10.1016/j.jpba.2015.06.016.
- 1047 [82] A.T. Caspar, F. Westphal, M.R. Meyer, H.H. Maurer, LC-high resolution-MS/MS for  
1048 identification of 69 metabolites of the new psychoactive substance 1-(4-ethylphenyl)-N-[(2-  
1049 methoxyphenyl)methyl] propane-2-amine (4-EA-NBOMe) in rat urine and human liver S9  
1050 incubates and comparison of its screening power wit, *Anal. Bioanal. Chem.* 410 (2018) 897–  
1051 912. doi:10.1007/s00216-017-0526-0.
- 1052 [83] P. Vervliet, O. Mortelé, C. Gys, M. Degreef, K. Lanckmans, K. Maudens, A. Covaci, A.L.N. van

- 1053 Nuijs, F.Y. Lai, Suspect and non-target screening workflows to investigate the in vitro and in  
1054 vivo metabolism of the synthetic cannabinoid 5CI-THJ-018, *Drug Test. Anal.* 11 (2019) 479–  
1055 491. doi:10.1002/dta.2508.
- 1056 [84] X. Diao, M.A. Huestis, New synthetic cannabinoids metabolism and strategies to best identify  
1057 optimal marker metabolites, *Front. Chem.* 7 (2019) 1–15. doi:10.3389/fchem.2019.00109.
- 1058 [85] Y. Djoumbou-Feunang, J. Fiamoncini, A. Gil-de-la-Fuente, R. Greiner, C. Manach, D.S. Wishart,  
1059 BioTransformer: a comprehensive computational tool for small molecule metabolism  
1060 prediction and metabolite identification, *J. Cheminform.* 11 (2019) 2. doi:10.1186/s13321-  
1061 018-0324-5.
- 1062 [86] S. Kern, K. Fenner, H.P. Singer, R.P. Schwarzenbach, J. Hollender, Identification of  
1063 Transformation Products of Organic Contaminants in Natural Waters by Computer-Aided  
1064 Prediction and High-Resolution Mass Spectrometry, *Environ. Sci. Technol.* 43 (2009) 7039–  
1065 7046. doi:10.1021/es901979h.
- 1066 [87] C. Hug, N. Ulrich, T. Schulze, W. Brack, M. Krauss, Identification of novel micropollutants in  
1067 wastewater by a combination of suspect and nontarget screening, *Environ. Pollut.* 184 (2014)  
1068 25–32. doi:10.1016/j.envpol.2013.07.048.
- 1069 [88] N.A. Alygizakis, S. Samanipour, J. Hollender, M. Ibáñez, S. Kaserzon, V. Kokkali, J.A. Van  
1070 Leerdam, J.F. Mueller, M. Pijnappels, M.J. Reid, E.L. Schymanski, J. Slobodnik, N.S. Thomaidis,  
1071 K. V. Thomas, Exploring the Potential of a Global Emerging Contaminant Early Warning  
1072 Network through the Use of Retrospective Suspect Screening with High-Resolution Mass  
1073 Spectrometry, *Environ. Sci. Technol.* 52 (2018) 5135–5144. doi:10.1021/acs.est.8b00365.
- 1074 [89] L. Bijlsma, E. Emke, F. Hernández, P. De Voogt, Performance of the linear ion trap Orbitrap  
1075 mass analyzer for qualitative and quantitative analysis of drugs of abuse and relevant  
1076 metabolites in sewage water, *Anal. Chim. Acta.* 768 (2013) 102–110.  
1077 doi:10.1016/j.aca.2013.01.010.
- 1078 [90] T. Reemtsma, U. Berger, H.P.H. Arp, H. Gallard, T.P. Knepper, M. Neumann, J.B. Quintana, P.  
1079 De Voogt, Mind the Gap: Persistent and Mobile Organic Compounds - Water Contaminants  
1080 That Slip Through, *Environ. Sci. Technol.* 50 (2016) 10308–10315.  
1081 doi:10.1021/acs.est.6b03338.
- 1082 [91] J.A. Baz-Lomba, M.J. Reid, K. V. Thomas, Target and suspect screening of psychoactive  
1083 substances in sewage-based samples by UHPLC-QTOF, *Anal. Chim. Acta.* 914 (2016) 81–90.  
1084 doi:10.1016/j.aca.2016.01.056.
- 1085 [92] Y. Picó, D. Barceló, Transformation products of emerging contaminants in the environment  
1086 and high-resolution mass spectrometry: A new horizon, *Anal. Bioanal. Chem.* 407 (2015)  
1087 6257–6273. doi:10.1007/s00216-015-8739-6.
- 1088 [93] Z. Qian, Z. Hua, C. Liu, W. Jia, Four types of cannabimimetic indazole and indole derivatives,  
1089 ADB-BINACA, AB-FUBICA, ADB-FUBICA, and AB-BICA, identified as new psychoactive  
1090 substances, *Forensic Toxicol.* 34 (2016) 133–143. doi:10.1007/s11419-015-0297-2.
- 1091 [94] I. González-Mariño, E. Gracia-Lor, R. Bagnati, C.P.B. Martins, E. Zuccato, S. Castiglioni,  
1092 Screening new psychoactive substances in urban wastewater using high resolution mass  
1093 spectrometry, *Anal. Bioanal. Chem.* 408 (2016) 4297–4309. doi:10.1007/s00216-016-9521-0.
- 1094 [95] H. Oberacher, K. Arnhard, Current status of non-targeted liquid chromatography-tandem  
1095 mass spectrometry in forensic toxicology, *TrAC Trends Anal. Chem.* 84 (2016) 94–105.  
1096 doi:10.1016/j.trac.2015.12.019.

- 1097 [96] X. Zhu, Y. Chen, R. Subramanian, Comparison of information-dependent acquisition, SWATH,  
1098 and MS All techniques in metabolite identification study employing ultrahigh-performance  
1099 liquid chromatography-quadrupole time-of-flight mass spectrometry, *Anal. Chem.* 86 (2014)  
1100 1202–1209. doi:10.1021/ac403385y.
- 1101 [97] A.T. Roemmelt, A.E. Steuer, T. Kraemer, Liquid Chromatography, in Combination with a  
1102 Quadrupole Time-of-Flight Instrument, with Sequential Window Acquisition of All Theoretical  
1103 Fragment-Ion Spectra Acquisition: Validated Quantification of 39 Antidepressants in Whole  
1104 Blood As Part of a Simultane, *Anal. Chem.* 87 (2015) 9294–9301.  
1105 doi:10.1021/acs.analchem.5b02031.
- 1106 [98] J. Náchter-Mestre, M. Ibáñez, R. Serrano, C. Boix, L. Bijlsma, B.T. Lunestad, R. Hannisdal, M.  
1107 Alm, F. Hernández, M.H.G. Berntssen, Investigation of pharmaceuticals in processed animal  
1108 by-products by liquid chromatography coupled to high-resolution mass spectrometry,  
1109 *Chemosphere.* 154 (2016) 231–239. doi:10.1016/j.chemosphere.2016.03.091.
- 1110 [99] R. Bade, J.M. White, C. Gerber, Qualitative and quantitative temporal analysis of licit and illicit  
1111 drugs in wastewater in Australia using liquid chromatography coupled to mass spectrometry,  
1112 *Anal. Bioanal. Chem.* 410 (2018) 529–542. doi:10.1007/s00216-017-0747-2.
- 1113 [100] R. Bade, B.J. Tschärke, J.M. White, S. Grant, J.F. Mueller, J. O'Brien, K. V. Thomas, C. Gerber,  
1114 LC-HRMS suspect screening to show spatial patterns of New Psychoactive Substances use in  
1115 Australia, *Sci. Total Environ.* 650 (2019) 2181–2187. doi:10.1016/j.scitotenv.2018.09.348.
- 1116 [101] A. Urbas, T. Schoenberger, C. Corbett, K. Lippa, F. Rudolphi, W. Robien, NPS Data Hub: A web-  
1117 based community driven analytical data repository for new psychoactive substances, *Forensic  
1118 Chem.* 9 (2018) 76–81. doi:10.1016/j.forc.2018.05.003.
- 1119 [102] M. Mardal, M.F. Andreasen, C.B. Møllerup, P. Stockham, R. Telving, N.S. Thomaidis, K.S.  
1120 Diamanti, K. Linnet, P.W. Dalsgaard, HighResNPS.com: An Online Crowd-Sourced HR-MS  
1121 Database for Suspect and Non-targeted Screening of New Psychoactive Substances, *J. Anal.  
1122 Toxicol.* 43 (2019) 520–527. doi:10.1093/jat/bkz030.
- 1123 [103] HighRes NPS Community, HighResNPS, HighResNPS. (n.d.). <http://highresnps.forensic.ku.dk/>.
- 1124 [104] J.Z. Seither, R. Hindle, L.E. Arroyo-Mora, A.P. DeCaprio, Systematic analysis of novel  
1125 psychoactive substances. I. Development of a compound database and HRMS spectral library,  
1126 *Forensic Chem.* 9 (2018) 12–20. doi:10.1016/j.forc.2018.03.003.
- 1127 [105] M. Mardal, S.S. Johansen, A.B. Davidsen, R. Telving, J.R. Jørnil, P.W. Dalsgaard, J.B.  
1128 Hasselstrøm, Å.M. Øiestad, K. Linnet, M.F. Andreasen, Postmortem analysis of three  
1129 methoxyacetylfentanyl-related deaths in Denmark and in vitro metabolite profiling in pooled  
1130 human hepatocytes, *Forensic Sci. Int.* 290 (2018) 310–317.  
1131 doi:10.1016/j.forsciint.2018.07.020.
- 1132 [106] H. Horai, M. Arita, S. Kanaya, Y. Nihei, T. Ikeda, K. Suwa, Y. Ojima, K. Tanaka, S. Tanaka, K.  
1133 Aoshima, Y. Oda, Y. Kakazu, M. Kusano, T. Tohge, F. Matsuda, Y. Sawada, M.Y. Hirai, H.  
1134 Nakanishi, K. Ikeda, N. Akimoto, T. Maoka, H. Takahashi, T. Ara, N. Sakurai, H. Suzuki, D.  
1135 Shibata, S. Neumann, T. Iida, K. Tanaka, K. Funatsu, F. Matsuura, T. Soga, R. Taguchi, K. Saito,  
1136 T. Nishioka, MassBank: A public repository for sharing mass spectral data for life sciences, *J.  
1137 Mass Spectrom.* 45 (2010) 703–714. doi:10.1002/jms.1777.
- 1138 [107] Z. Qian, W. Jia, T. Li, Z. Hua, C. Liu, Identification of five pyrrolidiny substituted cathinones  
1139 and the collision-induced dissociation of electrospray-generated pyrrolidiny substituted  
1140 cathinones, *Drug Test. Anal.* 9 (2017) 778–787. doi:10.1002/dta.2035.

- 1141 [108] M. Ibáñez, J.V. Sancho, L. Bijlsma, A.L.N. van Nuijs, A. Covaci, F. Hernández, Comprehensive  
1142 analytical strategies based on high-resolution time-of-flight mass spectrometry to identify  
1143 new psychoactive substances, *TrAC - Trends Anal. Chem.* 57 (2014) 107–117.  
1144 doi:10.1016/j.trac.2014.02.009.
- 1145 [109] D. Zuba, K. Sekuła, A. Buczek, Identification and characterization of 2,5-dimethoxy-4-nitro-β-  
1146 phenethylamine (2C-N) - A new member of 2C-series of designer drug, *Forensic Sci. Int.* 222  
1147 (2012) 298–305. doi:10.1016/j.forsciint.2012.07.006.
- 1148 [110] S.P. Elliott, S.D. Brandt, S. Freeman, R.P. Archer, AMT (3-(2-aminopropyl)indole) and 5-IT (5-  
1149 (2-aminopropyl)indole): An analytical challenge and implications for forensic analysis, *Drug*  
1150 *Test. Anal.* 5 (2013) 196–202. doi:10.1002/dta.1420.
- 1151 [111] R. Aalizadeh, M.-C. Nika, N.S. Thomaidis, Development and application of retention time  
1152 prediction models in the suspect and non-target screening of emerging contaminants, *J.*  
1153 *Hazard. Mater.* 363 (2019) 277–285. doi:10.1016/j.jhazmat.2018.09.047.
- 1154 [112] R. Bade, L. Bijlsma, T.H. Miller, L.P. Barron, J.V. Sancho, F. Hernández, Suspect screening of  
1155 large numbers of emerging contaminants in environmental waters using artificial neural  
1156 networks for chromatographic retention time prediction and high resolution mass  
1157 spectrometry data analysis, *Sci. Total Environ.* 538 (2015) 934–941.  
1158 doi:10.1016/j.scitotenv.2015.08.078.
- 1159 [113] S. Wolf, S. Schmidt, M. Müller-Hannemann, S. Neumann, In silico fragmentation for computer  
1160 assisted identification of metabolite mass spectra, *BMC Bioinformatics.* 11 (2010).  
1161 doi:10.1186/1471-2105-11-148.
- 1162 [114] C. Ruttkies, E.L. Schymanski, S. Wolf, J. Hollender, S. Neumann, MetFrag relaunched:  
1163 Incorporating strategies beyond in silico fragmentation, *J. Cheminform.* 8 (2016) 1–16.  
1164 doi:10.1186/s13321-016-0115-9.
- 1165 [115] J.W. Lee, Basics of ion mobility mass spectrometry, *Mass Spectrom. Lett.* 8 (2017) 79–89.  
1166 doi:10.5478/MSL.2017.8.4.79.
- 1167 [116] J. Regueiro, N. Negreira, M.H.G. Berntssen, Ion-mobility-derived collision cross section as an  
1168 additional identification point for multiresidue screening of pesticides in fish feed, *Anal.*  
1169 *Chem.* 88 (2016) 11169–11177. doi:10.1021/acs.analchem.6b03381.
- 1170 [117] L. Bijlsma, M.H.G. Berntssen, S. Merel, A Refined Nontarget Workflow for the Investigation of  
1171 Metabolites through the Prioritization by in Silico Prediction Tools, *Anal. Chem.* 91 (2019)  
1172 6321–6328. doi:10.1021/acs.analchem.9b01218.
- 1173 [118] A.S. Gelb, R.E. Jarratt, Y. Huang, E.D. Dodds, A study of calibrant selection in measurement of  
1174 carbohydrate and peptide ion-neutral collision cross sections by traveling wave ion mobility  
1175 spectrometry, *Anal. Chem.* 86 (2014) 11396–11402. doi:10.1021/ac503379e.
- 1176 [119] G. Paglia, J.P. Williams, L. Menikarachchi, J.W. Thompson, R. Tyldesley-Worster, S.  
1177 Halldórsson, O. Rolfsson, A. Moseley, D. Grant, J. Langridge, B.O. Palsson, G. Astarita, Ion  
1178 mobility derived collision cross sections to support metabolomics applications, *Anal. Chem.*  
1179 86 (2014) 3985–3993. doi:10.1021/ac500405x.
- 1180 [120] L. Fiebig, R. Laux, A collision cross section and exact ion mass database of the formulation  
1181 constituents polyethylene glycol 400 and polysorbate 80, *Int. J. Ion Mobil. Spectrom.* 19  
1182 (2016) 131–137. doi:10.1007/s12127-016-0195-2.
- 1183 [121] J. Hofmann, H.S. Hahm, P.H. Seeberger, K. Pagel, Identification of carbohydrate anomers  
1184 using ion mobility-mass spectrometry, *Nature.* 526 (2015) 241–244.

- 1185 doi:10.1038/nature15388.
- 1186 [122] J. Regueiro, A. Giri, T. Wenzl, Optimization of a Differential Ion Mobility Spectrometry-  
1187 Tandem Mass Spectrometry Method for High-Throughput Analysis of Nicotine and Related  
1188 Compounds: Application to Electronic Cigarette Refill Liquids, *Anal. Chem.* 88 (2016) 6500–  
1189 6508. doi:10.1021/acs.analchem.6b01241.
- 1190 [123] L. Bijlsma, R. Bade, A. Celma, L. Mullin, G. Cleland, S. Stead, F. Hernandez, J.V.J.V. Sancho,  
1191 Prediction of Collision Cross-Section Values for Small Molecules: Application to Pesticide  
1192 Residue Analysis, *Anal. Chem.* 89 (2017) 6583–6589. doi:10.1021/acs.analchem.7b00741.
- 1193 [124] L. Bijlsma, J. V. Sancho, F. Hernández, W.M.A. Niessen, Fragmentation pathways of drugs of  
1194 abuse and their metabolites based on QTOF MS/MS and MS E accurate-mass spectra, *J. Mass  
1195 Spectrom.* 46 (2011) 865–875. doi:10.1002/jms.1963.
- 1196 [125] R. Lian, F. Zhang, Y. Zhang, Z. Wu, H. Ye, C. Ni, X. Lv, Y. Guo, Ion mobility derived collision  
1197 cross section as an additional measure to support the rapid analysis of abused drugs and toxic  
1198 compounds using electrospray ion mobility time-of-flight mass spectrometry, *Anal. Methods.*  
1199 10 (2018) 749–756. doi:10.1039/C7AY02808C.
- 1200 [126] C. Tejada-Casado, M. Hernández-Mesa, F. Monteau, F.J. Lara, M. del Olmo-Iruela, A.M.  
1201 García-Campaña, B. Le Bizec, G. Dervilly-Pinel, Collision cross section (CCS) as a  
1202 complementary parameter to characterize human and veterinary drugs, *Anal. Chim. Acta.*  
1203 1043 (2018) 52–63. doi:10.1016/j.aca.2018.09.065.
- 1204 [127] C.B. Mollerup, M. Mardal, P.W. Dalsgaard, K. Linnet, L.P. Barron, Prediction of collision cross  
1205 section and retention time for broad scope screening in gradient reversed-phase liquid  
1206 chromatography-ion mobility-high resolution accurate mass spectrometry, *J. Chromatogr. A.*  
1207 1542 (2018) 82–88. doi:10.1016/j.chroma.2018.02.025.
- 1208 [128] Z. Zhou, X. Shen, J. Tu, Z.J. Zhu, Large-scale prediction of collision cross-section values for  
1209 metabolites in ion mobility-mass spectrometry, *Anal. Chem.* 88 (2016) 11084–11091.  
1210 doi:10.1021/acs.analchem.6b03091.
- 1211 [129] E. Gracia-Lor, S. Castiglioni, R. Bade, F. Been, E. Castrignanò, A. Covaci, I. González-Mariño, E.  
1212 Hapeshi, B. Kasprzyk-Hordern, J. Kinyua, F.Y. Lai, T. Letzel, L. Lopardo, M.R. Meyer, J. O'Brien,  
1213 P. Ramin, N.I. Rousis, A. Rydevik, Y. Ryu, M.M. Santos, I. Senta, N.S. Thomaidis, S. Veloutsou,  
1214 Z. Yang, E. Zuccato, L. Bijlsma, Measuring biomarkers in wastewater as a new source of  
1215 epidemiological information: Current state and future perspectives, *Environ. Int.* 99 (2017)  
1216 131–150. doi:10.1016/j.envint.2016.12.016.
- 1217 [130] E. Gracia-Lor, E. Zuccato, S. Castiglioni, Refining correction factors for back-calculation of illicit  
1218 drug use, *Sci. Total Environ.* 573 (2016) 1648–1659. doi:10.1016/j.scitotenv.2016.09.179.
- 1219 [131] L. Bijlsma, R. Gil-Solsona, F. Hernández, J.V. Sancho, What about the herb? A new  
1220 metabolomics approach for synthetic cannabinoid drug testing, *Anal. Bioanal. Chem.* 410  
1221 (2018) 5107–5112. doi:10.1007/s00216-018-1182-8.
- 1222 [132] A.T. Caspar, S.D. Brandt, A.E. Stoeber, M.R. Meyer, H.H. Maurer, Metabolic fate and  
1223 detectability of the new psychoactive substances 2-(4-bromo-2,5-dimethoxyphenyl)- N- [(2-  
1224 methoxyphenyl)methyl]ethanamine (25B-NBOMe) and 2-(4-chloro-2,5-dimethoxyphenyl)- N-  
1225 [(2-methoxyphenyl)methyl]ethanamine (25C-NBOMe) in human and r, *J. Pharm. Biomed.*  
1226 *Anal.* 134 (2017) 158–169. doi:10.1016/j.jpba.2016.11.040.
- 1227 [133] E.D. Amato, A. Covaci, R.M. Town, J. Hereijgers, B. Bellekens, V. Giacometti, T. Breugelmans,  
1228 M. Weyn, F. Dardenne, L. Bervoets, R. Blust, A novel active-passive sampling approach for

- 1229 measuring time-averaged concentrations of pollutants in water, *Chemosphere*. 209 (2018)  
1230 363–372. doi:10.1016/j.chemosphere.2018.06.079.
- 1231 [134] Y. Zhang, T. Zhang, C. Guo, S. Hou, Z. Hua, J. Lv, Y. Zhang, J. Xu, Development and application  
1232 of the diffusive gradients in thin films technique for simultaneous measurement of  
1233 methcathinone and ephedrine in surface river water, *Sci. Total Environ.* 618 (2018) 284–290.  
1234 doi:10.1016/j.scitotenv.2017.11.068.
- 1235 [135] A. Krueve, Semi-quantitative non-target analysis of water with liquid chromatography/high-  
1236 resolution mass spectrometry: How far are we?, *Rapid Commun. Mass Spectrom.* 33 (2019)  
1237 54–63. doi:10.1002/rcm.8208.
- 1238 [136] R. Bade, J.M. White, L. Nguyen, B.J. Tschärke, J.F. Mueller, J.W. O’Brien, K. V. Thomas, C.  
1239 Gerber, Determining changes in new psychoactive substance use in Australia by wastewater  
1240 analysis, *Sci. Total Environ.* 731 (2020) 139209. doi:10.1016/j.scitotenv.2020.139209.
- 1241
- 1242

1243 **Figure captions**

1244 **Figure 1:** Identification of two co-eluting NPS, butyryl fentanyl ( $m/z$  351.2431) and furanylfentanyl  
1245 ( $m/z$  375.2067) in a spiked wastewater sample using Sequential Window Acquisition of all  
1246 Theoretical fragment ion spectra (SWATH). (A) overlapping extraction ion chromatograms  
1247 (XICs) of the two NPS with chromatographic peaks eluting at 12.50 min (top); full scan  
1248 acquisition mass spectra with low collision energy (LE) (10 V) at retention time  $12.50 \pm 0.10$   
1249 min (bottom). (B) SWATH mass window  $m/z$  340.2-357.4, XIC at  $m/z$  351.24 (middle) and high  
1250 collision energy (HE) mass spectra (bottom); (C) SWATH mass window  $m/z$  372.6-389.8, XIC at  
1251  $m/z$  375.21 (middle) and HE mass spectra (bottom).

1252 **Figure 2:** Tentative identification of 4'-chloro- $\alpha$ -pyrrolidinopropiophenone (4-chloro- $\alpha$ -PPP) in a  
1253 pooled urine sample. (A) Extracted ion chromatogram of 4-chloro- $\alpha$ -PPP and  $^{37}\text{Cl}$  isotope (top);  
1254 Product ion mass spectra of  $[\text{M}+\text{H}]^+$  at  $m/z$  238.10 (bottom). (B) Structure of 4-chloro- $\alpha$ -PPP  
1255 (top); Zoom in the range of fragment ion with  $m/z$  167 at resolution (R) of 35.000 Full Width at  
1256 Half Maximum (FWHM) (bottom).

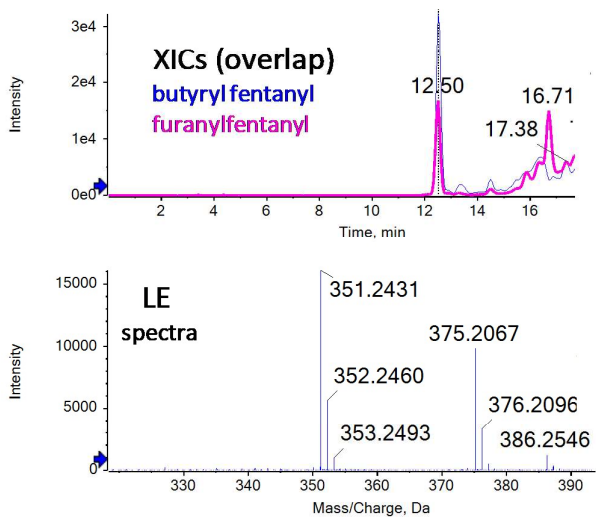
1257 **Figure 3:** Identification of  $\alpha$ -methyltryptamine in a raw wastewater sample using QTOF MS. (A) feature  
1258 detection of  $m/z$  175.1235 at 4.51 min (top, insert) together with the low collision energy (LE)  
1259 spectra (top) and high collision energy (HE) spectra with emphasis on  $m/z$  130-145 (grey areas)  
1260 (bottom); (B) Structure, fragment ions, LE and HE spectra of  $\alpha$ -methyltryptamine; (C) Structure,  
1261 fragment ions, LE and HE spectra of 5-(2-aminopropyl)indole.

1262 **Figure 4:** Identification of ketamine in a raw wastewater sample using IMS QTOF MS. (A) feature  
1263 detection of  $m/z$  238.0993 at 3.33 min and drift time (DT) 4.89 ms, yellow arrow (top) ( $\approx 70$   
1264  $\mu\text{s}/\text{scan}$ ); co-eluting ions at 3.33 min illustrated as red or black dots and separated by DT. Blue  
1265 highlighted areas are the DT ranges of  $4.89 \pm 0.20$  ms at  $m/z$  238.0993 at low collision energy  
1266 (LE) and high collision energy (HE) (bottom). (B) LE and HE mass spectra without IMS DT  
1267 alignment (top); LE and HE mass spectra with IMS DT alignment (bottom).

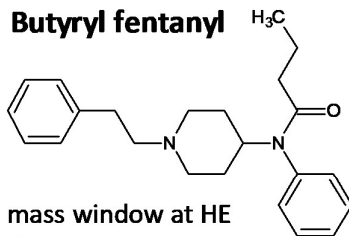


Full scan at LE

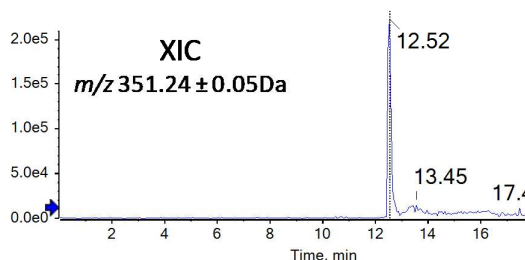
$m/z$  351.24  $\pm$  0.05Da XIC butyryl fentanyl  
 $m/z$  375.21  $\pm$  0.05Da XIC furanylfentanyl



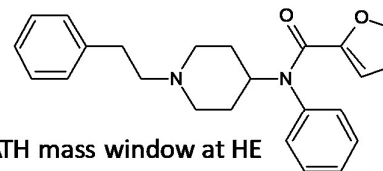
Butyryl fentanyl



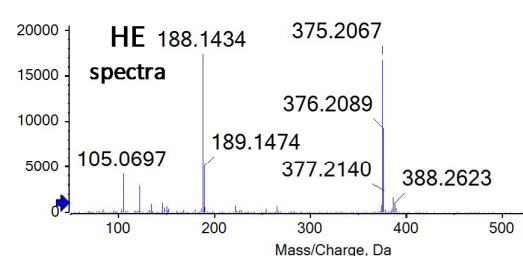
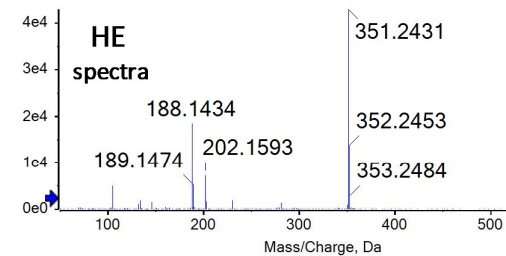
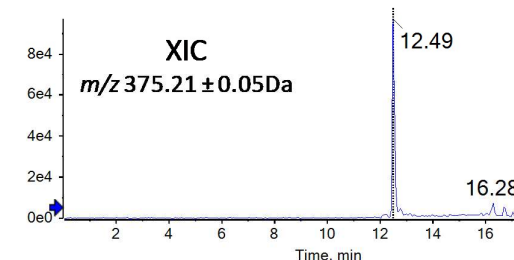
SWATH mass window at HE  
 $m/z$  340.2-357.4



Furanylfentanyl



SWATH mass window at HE  
 $m/z$  372.6-389.8



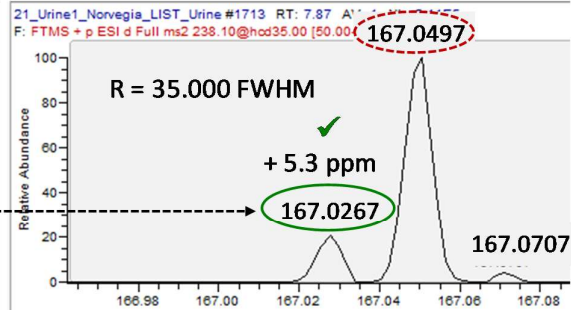
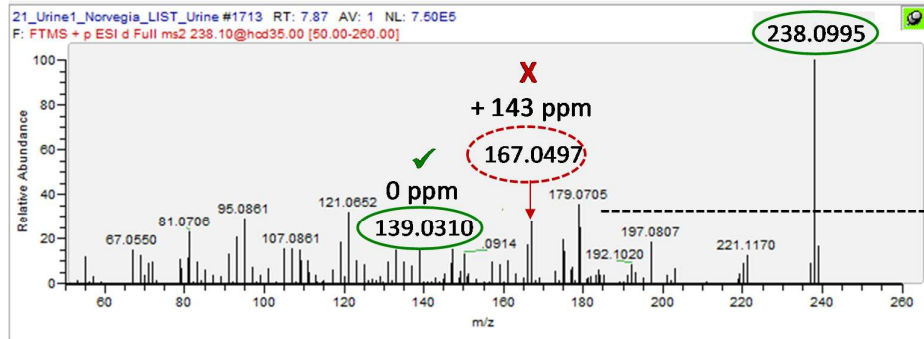
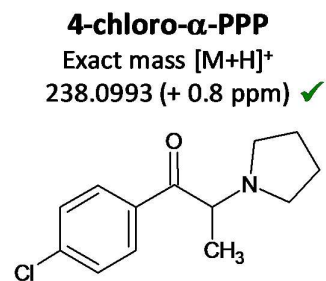
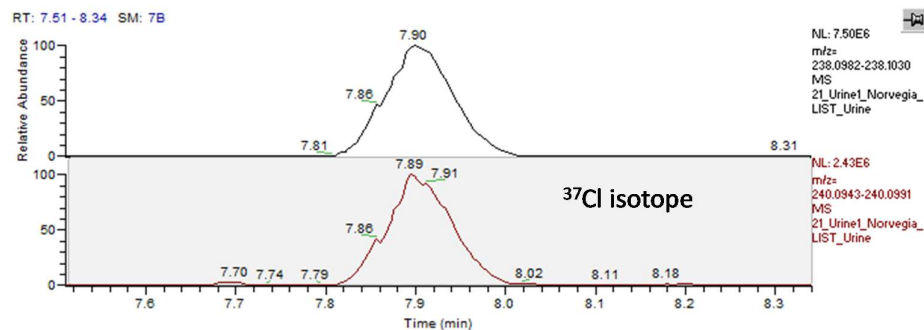
A

B

C

1268

1269 Figure 1

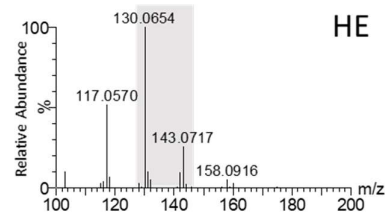
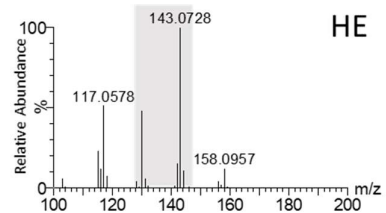
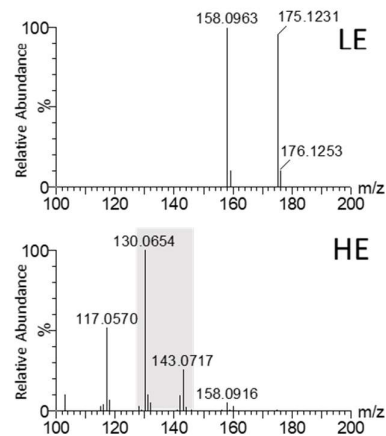
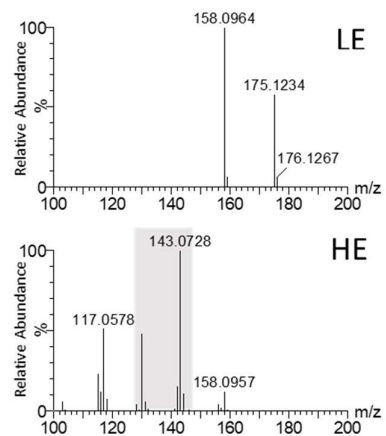
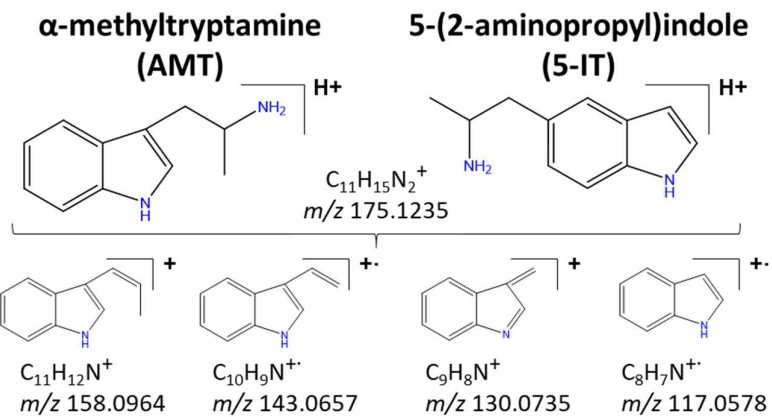
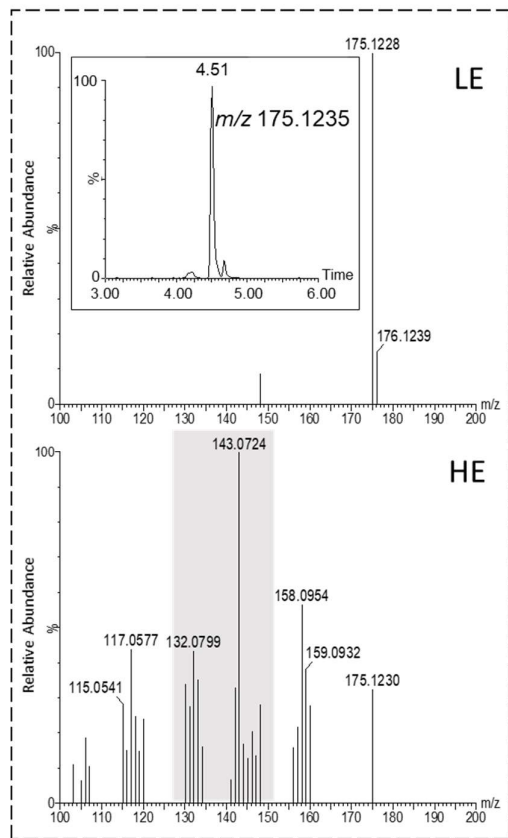


A

B

1270

1271 Figure 2



A

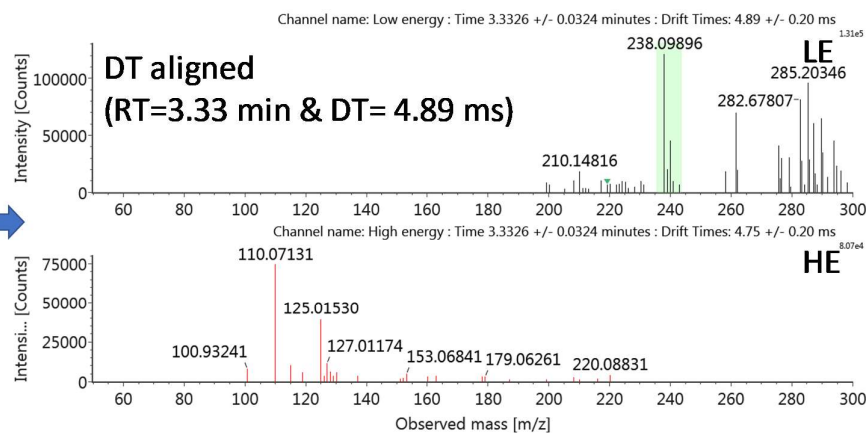
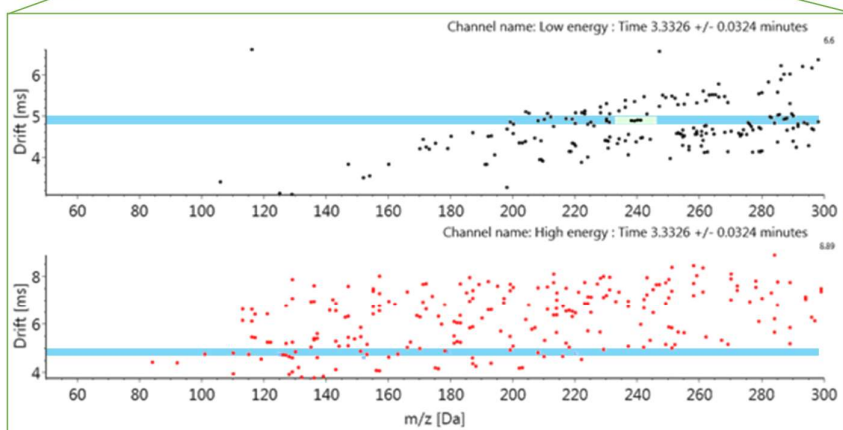
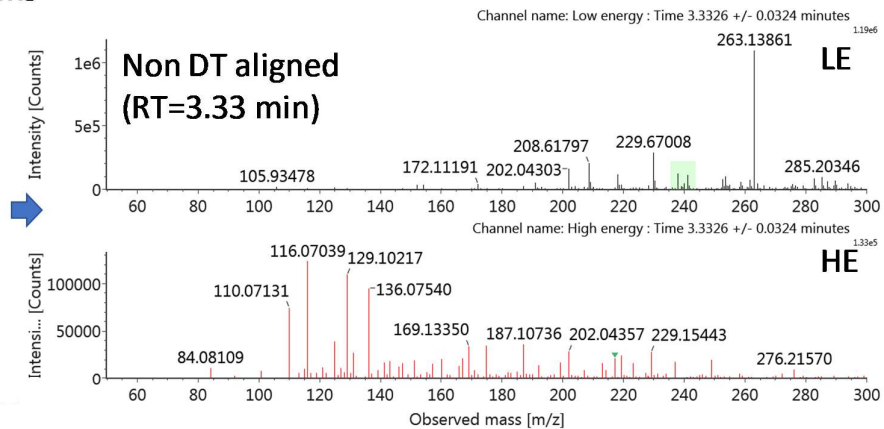
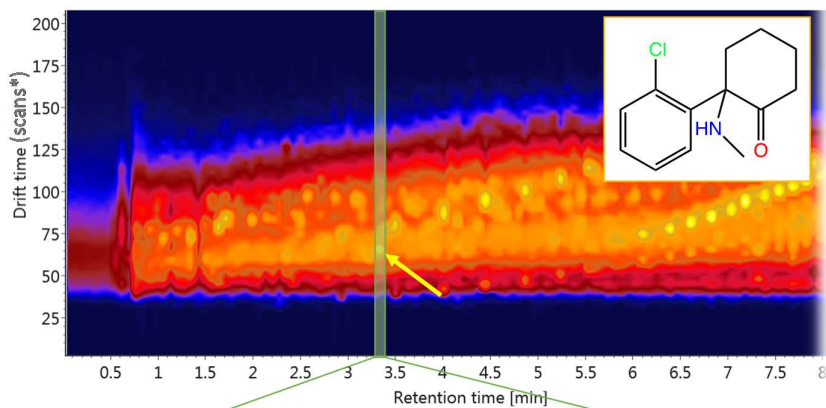
B

C

1272

1273 **Figure 3**

Ketamine, RT=3.33 min,  $m/z$  ( $[M+H]^+$ )= 238.0993, DT= 4.89 ms



A

B

1274

1275 **Figure 4**