1	Simultaneous determination of dechloranes, polybrominated diphenyl ethers and novel													
2	brominated flame retardants in food and serum													
3														
4	Carlos Sales ^{a*} , Giulia Poma ^b , Govindan Malarvannan ^b , Tania Portolés ^a ,													
5	Joaquin Beltrán ^ª , Adrian Covaci ^{b*}													
6														
7	^a Research Institute for Pesticides and Water, University Jaume I, E-12071 Castellón, Spain													
8	^b Toxicological Centre, Department of Pharmaceutical Sciences, University of Antwerp,													
9	Universiteitsplein 1, B-2610 Wilrijk, Belgium													
10														
11	*Tel: 0034964387391, Fax: 0034964387368, E-mail: <u>casales@uji.es</u>													
12	*Tel: 032652498, Fax: 032652722, E-mail: <u>adrian.covaci@uantwerpen.be</u>													

14 Abstract

15 A sensitive method for the simultaneous quantification of dechloranes, polybrominated diphenyl 16 ethers (PBDEs), and novel brominated flame retardants (NBFRs) has been developed for gas 17 chromatography (GC) coupled to tandem mass spectrometry operating in electron capture negative 18 ionization (ECNI) mode. The major advance has been achieved by combining selected ion monitoring 19 (SIM) and multiple reaction monitoring (MRM) modes in well-defined time windows, to determine dechloranes, PBDEs and NBFRs at pg g⁻¹ level in one single analysis in complex matrix biological 20 21 samples. From the chromatographic point of view, efforts were devoted to study several injection 22 modes using multimode inlet (MMI) in order to obtain low instrumental detection limits, necessary for 23 trace compounds such as Dechlorane Plus (DP) isomers. Method performance was also evaluated: calibration curves were linear from 20 fg μ L⁻¹ to 100 pg μ L⁻¹ for the studied compounds, with method 24 detection limits at levels of 50 fg g⁻¹ for DPs. Repeatability and reproducibility, expressed as relative 25 26 standard deviation, were better than 5% even in solvent vent mode for the injection of standards. The 27 application to a wide range of complex samples (including food, human, and animal serum samples) indicated a sensitive and reliable way to quantify at the pg g^{-1} level four HNs, Dechlorane Plus (*anti*-DP 28 29 and syn-DP) and two of their homologues (Dechlorane-602 and Dechlorane-603), 11 PBDE congeners 30 (no. 28, 47, 49, 66, 85, 99, 100, 153, 154, 183, and 209), and five novel BFRs, *i.e.*, decabromodiphenyl 31 ethane (DBDPE), 1,2-bis(2,4,6-tribromo-phenoxy)ethane (BTBPE), hexabromobenzene (HBB), 2,3,4,5-32 tetrabromo-ethylhexyl-benzoate (TBB) and tetrabromophthalate (TBPH).

33

34 Keywords

Chemical ionization, gas chromatography, brominated flame retardants, Dechloranes, large volume
 injection, biological matrices

38 **1. Introduction**

39 Halogenated flame retardants (HFRs), including chlorinated and brominated compounds, are used to 40 prevent ignition and combustion of flammable materials, widely employed in furniture, plastics, foams, 41 and textile upholstery, among other products [1]. HFRs have been detected in various environmental 42 and food samples as they are released into air, soil and water due to manufacture, improper handling, 43 and disposal of HFR-containing products and materials [2]. Among them, polybrominated diphenyl 44 ethers (PBDEs) have been extensively investigated, as a consequence of their past usage, toxicity and 45 persistence in the food chain [3, 4]. As a result of bans applied to commercial PBDE mixtures, there is 46 an increasing production and use of alternative HFRs [5, 6]. Nevertheless, apart from monitoring these 47 HFRs, the determination of PBDEs is still necessary for monitoring purposes and to assess their 48 replacement efficiency [7, 8]. A scheme displaying the different structures of the investigated HFRs is 49 shown in Figure 1.

50 There is a large amount of literature regarding the analysis of PBDEs and novel brominated flame 51 retardants (NBFRs) by gas chromatography-mass spectrometry (GC-MS) and GC-MS/MS using electron 52 capture negative ionization (ECNI) and electron ionization (EI) [9]. More recently, atmospheric 53 pressure chemical ionization (APCI) [10, 11] has also been reported for the analysis of brominated FRs. 54 Both ECNI and APCI offer good sensitivity when compared to EI, while the specificity when using APCI 55 and EI in MS/MS experiments is higher than the obtained by monitoring bromide ions in ECNI [10]. For 56 the determination of Dechloranes, the majority of studies performed so far used EI-MS(/MS), with 57 insufficient detection limits in some cases [12, 13] or ECNI-MS(/MS) with the need of an additional 58 injection, separated from PBDEs [14–16]. GC-EI-MS/MS methods monitor transitions derived from the 59 molecular ion to m/z 237 and m/z 228 [16]. Similar to PBDEs, the analysis of DPs can benefit of selecting 60 more specific transitions coming from the molecular ion by using softer ionization sources. DP isomers 61 constitute a special case study, as they have quite a particular fragmentation behaviour. Several 62 studies have investigated the different fragmentation of anti- and syn- DP isomers under variable ECNI source temperatures, either in full-scan [17] or in SIM experiments [15], but not yet in MRM
experiments.

Human biomonitoring data on dechloranes is not extensive, but there are few articles on studies
investigating their presence and levels in serum, e.g. from China, where the highest levels have been
reported near e-waste recycling plants [18–20], and from Canada [21], Norway [22, 23] or Germany
[24]. In all these studies, the limits of detection for DPs were in the pg g⁻¹ level.

Against this background, the availability of a method with the benefits of sensitivity and specificity for Dechloranes and sensitivity for PBDEs and other flame retardants in a single analysis using a chemical ionization (CI) source in negative mode could be beneficial for monitoring laboratories. The aim of this work was the development of a methodology for the simultaneous analysis of HFRs of high concern at low pg g⁻¹ levels in a wide range of complex samples, such as food, human and animal serum samples. Such improvement of the analytical methods will be useful in the currently running biomonitoring schemes, such as the Flemish Environment and Health study.

76

77 **2.** Materials and methods

78 **2.1. Chemicals and reagents**

79 Standards of BDE-28, -49, -47, -99, -100, -153, -154, -138, -183, and -209, 1,2-Bis(2,4,6-80 tribromophenoxy)ethane (BTBPE), syn-DP and anti-DP isomers, 2-ethylhexyl-2,3,4,5 tetrabromo-81 benzoate (TBB), 2,3,4,5-tetrabromophthalate (TBPH), hexabromobenzene (HBB), dechlorane-602 (Dec-602), dechlorane-603 (Dec-603), isotopically labelled internal standards (IS) ¹³C-BDE-209, ¹³C-82 TBPH, ¹³C-TBB, ¹³C-syn-DP, and ¹³C-anti-DP were purchased from Wellington Laboratories (Guelph, ON, 83 84 Canada). Recovery standard (RS) CB-207 was purchased from Dr. Ehrenstorfer Laboratories (Augsburg, 85 Germany). Polypropylene (PP) tubes (15 mL) were obtained from Greiner Bio-one (Belgium). Empty PP 86 cartridges (25 mL) were purchased from Grace (Lokeren, Belgium), while Florisil® cartridges (500 mg, 87 3 mL) and empty PP cartridges (6 mL) were purchased from Supelco (Bellefonte, PA, USA). Silica gel, 88 anhydrous sodium sulphate (Na₂SO₄) and concentrated sulfuric acid (H₂SO₄, 98%) were purchased from Merck (Darmstadt, Germany). All solvents were of chromatography grade: n-hexane was purchased
from Acros Organics (Belgium); dichloromethane (DCM), iso-octane, toluene and acetonitrile (ACN)
were purchased from Merck.

92

93 **2.2 Sample Treatment**

Food samples (including biscuits, smoked salmon, and chicken eggs) were treated as indicated in a previous work [25]. Briefly, samples were homogenized, freeze-dried, and stored at -20 °C until analysis. The samples were weighted in pre-washed 15 mL polypropylene (PP) tubes, and spiked with the IS mixture. After spiking, samples were extracted by solid-liquid extraction (SLE) with ACN:toluene (9:1, v/v). After a two-step clean-up (performed on Florisil® and acidified silica 5%), the samples were evaporated to dryness and reconstituted in 100 µL of the recovery standard (RS) (CB-207 in isooctane:toluene; 9:1, v/v) and transferred to amber injection vials for GC-ECNI-MS(/MS) analysis.

Serum samples including hyena, cheetah and lion (Zoo Antwerp, Belgium), sea eagle plasma (Trondheim, Norway) and human cord blood (Flemish Environment and Health study II – Flemish newborns) were extracted according to the method described elsewhere [26], with slight modifications. Solid-phase extraction (SPE) on OASIS HLB cartridges was used followed by clean-up on 1 g of acidified silica 44% and eluted with 10 mL n-hexane:dichloromethane (1:1, v/v). The cleaned extract was evaporated to incipient dryness and re-dissolved in 100 µL iso-octane.

107

108 **2.3 GC-(ECNI)-MS(/MS)**

109 The chromatographic analysis was performed using an Agilent 7890B gas chromatograph, equipped 110 with an Agilent 7693A autosampler with Multimode Inlet (MMI), coupled to a triple quadrupole mass 111 spectrometer, 7000C (Agilent Technologies Inc., Palo Alto, CA, USA), with a CI source working in 112 electron capture negative ionization mode. Methane was used as reagent gas at a flow of 2 mL min⁻¹. 113 The GC separation was performed using a fused silica a ZB-semivolatiles capillary column (5% phenylarylene-95% dimethyl-polysiloxane) with a length of 20 m x 0.18 mm ID and a film thickness of 0.18 μ m (Phenomenex, Torrance, CA, USA) working at a ramped flow from 1 mL min⁻¹ (14 min) with 10 mL min⁻¹ to 2 mL min⁻¹ (10.9 min) of helium (99.999 %; Air Liquide, Liège, Belgium). The oven program was set as follows: 90 °C (1.25 min); 30 °C min⁻¹ to 240 °C; then 10 °C min⁻¹ to 325 °C, stay 10.4 min with a total run time of 25 min. The injection of 2 μ L of sample extracts was performed in cold pulsed splitless mode with at a temperature of 80 °C and a pulse time of 1.25 min. The pulse pressure was set to 50.0 psi, with a split purge flow of 50 mL min⁻¹ and purge time of 1.25 min.

121

122 **3.** Results and discussion

123 **3.1. MS optimization**

124 Optimal m/z values for SIM of each compound were selected according to [27], while the optimal MRM 125 transitions for DPs were taken from reference [28], also considering the common fragmentation 126 pattern for every compound, usually leading to bromide ions. To achieve maximum sensitivity, 127 different collision energies were tested to study the fragmentation of syn- and anti-DPs in the collision 128 cell. Two ions from the isotopic pattern corresponding to M⁻⁺ (M+4 and M+6) were selected in the first 129 quadrupole and fragmentation was performed using a range of collision energies between 5 eV and 35 eV. A collision energy of 5 eV was optimal for the ¹³C-labelled DPs, while 10 eV was selected for the 130 131 native syn- and anti- DPs. Accordingly, the selected transitions were $654 \rightarrow 35$; $654 \rightarrow 37$ and $652 \rightarrow 35$ 132 corresponding to the fragmentation of the precursor m/z ions $[M+6]^{-1}$ and $[M+4]^{-1}$ for the native DPs 133 and $664 \rightarrow 35$ and $664 \rightarrow 37$ taking [M+6]⁻⁻ m/z ion as precursor for the ¹³C- DPs.

The source temperature was also optimized pursuing the maximum response for every analyte. Previous studies, [17] and [15], demonstrated that low source temperatures favour the detection of the molecular ion cluster, while higher temperatures (250 °C) had different effects on both isomers. According to De la Torre et al [17], a temperature of 150 °C provided similar spectra for both isomers, with the most abundant cluster being the one corresponding to the molecular ion [M]-. However, at 250 °C, the two isomers showed a different pattern. In the case of *syn*-DP, the cluster corresponding to the ion [M-6Cl]⁻ became the most abundant. Summarizing, higher temperature source provided
 more energy, hence favouring the dissociative electron capture process and increasing the abundance
 of fragment ions, whereas lower temperatures enhanced molecular ion abundance.

143 Accordingly, after obtaining low collision energies as the optimal for the determination of DPs, we 144 theorized that low source temperatures might enhance the formation in the ion source of the parent 145 ions for the DPs transitions. Hence, source temperatures of 250 °C, 225 °C, and 200 °C were tested. An 146 increase in the response of DPs was seen at lower source temperatures (Figure 2), while too low 147 temperatures could affect the sensitivity for PBDEs, for which detection relies in the fragmentation to 148 the bromide ion m/z 79. The temperature of 200 °C was hence chosen as a compromise for these 149 experiments. Selected quantification and qualification transitions and ions for each analyte are 150 summarized in Table 1. These findings add to the previous studies on the behaviour of DPs at different 151 source temperatures, as in MRM experiments, the formation of an abundant molecular pattern to be 152 selected as a parent ion, has been proved more sensitive than a high in-source fragmentation, which 153 leads to larger losses of chlorine atoms before entering in the first quadrupole.

154

155 **3.2. Analytical parameters**

156 To maximize the signal obtained for each analyte, the use of the multimode inlet in large volume 157 injection mode was considered. The possibility of starting at a low inlet temperature allowed the 158 injection of a higher volume of extract. Therefore, several injection configurations were tested: cold 159 pulsed splitless (2 μ L), and solvent vent (5 μ L, 2 x 5 μ L and 3 x 5 μ L). Figure 3 highlights the response 160 enhancement for the DP congeners when working at the three selected working conditions. Although 161 solvent vent injections enhance the sensitivity for DPs as well as for the rest of the selected 162 compounds, reproducibility and overloading issues were noticed when injecting extracts from fatty 163 matrices, so the injection of 2 µL in cold pulsed splitless mode was selected as optimal. To test the 164 reliability of the method, the repeatability of absolute area was studied in five repeated injections of standards at five different levels (20 fg μ L⁻¹, 100 fg μ L⁻¹, 1 pg μ L⁻¹, 20 pg μ L⁻¹ and 100 pg μ L⁻¹). The relative 165

166 standard deviation was below 5%. Linearity of the relative response of the different compounds (to 167 their ¹³C isotopically labelled or BDE internal standards) was studied by analyzing standard solutions, in triplicate (five levels), in the range of 20 fg μL^{-1} to 100 pg μL^{-1} . The correlation coefficients (r²) were 168 169 higher than 0.99 for every compound, with residuals lower than 2%. Special attention has to be paid 170 to the method sensitivity for DPs, which can be derived from **Figure 2** (injection of a 16 fg μ L⁻¹standard 171 solution in isooctane). Instrumental limits of detection (iLODs) were calculated as the lowest 172 concentration level giving a signal-to-noise ratio (S/N) of 3. These iLODs were determined to be around 1 fg μ L⁻¹ for syn-DP and 0.5 fg μ L⁻¹ for anti-DP, when injecting 2 μ L in cold pulsed splitless mode. The 173 174 iLODs were even lower when using solvent vent mode, as can be seen in Figure 3. Obtained iLODs are 175 summarized in Table 1. LODs and LOQs in real samples were estimated using the same criteria, by 176 extrapolation from the lowest responses (detectable and quantifiable) of every compound within the 177 analysed samples. These results are relevant especially for DP isomers, as their LODs and LOQs have 178 been lowered sensibly in comparison to previous studies. Table 2 lists the majority of previous studies 179 performed to detect and quantify DP isomers, indicating the systems used and the achieved 180 performance in each case in terms of LOD and LOQ.

181

3.3. Analysis of real samples

The enhanced capabilities of the presented method were finally tested using extracts of samples of food and human and animal serum previously analysed by GC-ECNI-MS, according to the method used for routine analysis and described elsewhere [25]. The developed methodology allowed the determination of trace quantities (below pg g⁻¹ range) of the selected PBDEs in several samples. In these samples, NBFRs could also be evidenced. A good agreement was found when comparing the quantification results of the new methodology with those given by the validated reference method [25] (at the levels achievable by the reference method).

Special emphasis was made on the capability of the methodology to detect DPs in most of analysed
samples. Due to the presence of these compounds in the procedural blanks, only the samples with DPs

192 relative area higher than 10 times their corresponding relative area in the blank were quantified. The 193 most remarkable results to highlight are: a pool of four cord blood human serum samples with 0.13 and 0.19 pg g^{-1} of syn-DP and anti-DP, respectively; a sea eagle plasma sample with 1.95 pg g^{-1} of syn-194 DP and 26 pg g⁻¹ of *anti*-DP, a chicken egg with 9 ng g⁻¹ of *syn*-DP and 29 ng g⁻¹ of *anti*-DP and a hyena 195 196 serum sample with 0.33 pg g⁻¹ of syn-DP. Dec-603 and Dec-602 were also quantified in human/animal serum ranging from 5 to 66 pg g⁻¹. Chromatograms with the quantification transition of DP isomers in 197 198 the mentioned samples can be seen in Figure 4 (4A for a procedure blank, biscuits, smoked salmon, 199 chicken egg and hyena extracts, and **4B** showing a cheetah serum, human cord blood (pool), sea eagle 200 serum, and two chicken egg extracts). Table 3 summarizes the concentration found for each analyte 201 in the samples.

The most contaminated samples corresponded, as expected, to captive animals from the Antwerp Zoo and the eggs of wild birds. It is also important to consider the differences found in the f-anti value. *Anti-DP* has been found to degrade faster than *syn-DP* at high temperatures and at e-waste sites [18], so the differences measured with this methodology, for example in the hyena sample, could help to assess for the degradation of these compounds in areas close to recycling facilities and monitor theirpresence of them in animals and humans.

208

209 4. Conclusions

210 The use of a method combining SIM and MRM acquisition modes in an ECNI source has demonstrated 211 high sensitivity for a wide range of HFRs, specifically for DP isomers, which have been detected in most 212 of analyzed samples, including procedural blanks. This combination of acquisition modes together with large volume injections allowed decreasing the LODs for DPs to fg g⁻¹ levels, which constitutes a 213 214 significant advancement compared to previous methodologies monitoring the molecular ion in SIM 215 mode or less sensitive transitions in EI-MS/MS. Nevertheless, the use of large volume injections can 216 be an issue for some fatty matrices and has to be carefully applied to selected samples. The method was applied to a wide range of complex matrices and was able to quantify DP isomers at low pg g^{-1} 217

- 218 levels in serum samples. This methodology is an important tool for the determination of HFRs at ultra-
- trace levels in food and biological samples, helping to monitor the release and occurrence of PBDEs,
- 220 HNs and NBFRs in the environment.
- 221
- 222

223 Acknowledgments

- The authors acknowledge the financial support of Universitat Jaume I (UJI-A2016-01) and Generalitat
 Valenciana, (research group of excellence PROMETEO/2009/054 and PROMETEO II 2014/023 and
 Collaborative Research on Environment and Food-Safety (ISIC/2012/016)). Carlos Sales acknowledges
 the COST Action ES1307 for the STSM grant which made possible his stay at the Toxicological Centre
 of Antwerp. Dr. Giulia Poma and Dr. Govindan Malarvannan acknowledge the University of Antwerp
 for their postdoctoral fellowships.
 Conflict of interest
- ----
- 232 The authors declare that they have no conflict of interest.
- 233

234 Figure Captions

- **Fig. 1** Scheme of the structures of the main compounds selected for the study.
- 236 Fig. 2 Variation in the peak area for the most sensitive MRM transition for DPs, for the injection of a
- 237 $\,$ standard mixture at 16 fg $\mu L^{\text{-1}}$ in isooctane at different source temperatures (200 °C, 225 °C and 250 $\,$

238 °C).

- 239 Fig. 3 Graphical comparison of the methodology performance for the injection of a DP mixture (16 fg
- 240 μL^{-1}). S/N = signal to noise ratio.
- Fig. 4 Chromatograms corresponding to the quantification transition of DPs for the injection of (A)
- procedural blank, biscuits, smoked salmon, chicken egg and hyena serum extracts, and (B) cheetah
- serum, human cord blood serum, chicken egg and sea eagle serum extracts.

245 References

- de Wit CA (2002) An overview of brominated flame retardants in the environment.
 Chemosphere 46:583–624.
- 248
 2. Watanabe I, Sakai S (2003) Environmental release and behavior of brominated flame
 249 retardants. Environ Int 29:665–682.
- Law RJ, Covaci A, Harrad S, Herzke D, Abdallah M a E, Fernie K, Toms LM, Takigami H (2014)
 Levels and trends of PBDEs and HBCDs in the global environment: Status at the end of 2012.
 Environ Int 65:147–158.
- 2534.Wang D, Li QX (2010) Application of mass spectrometry in the analysis of polybrominated254diphenyl ethers. Mass Spectrom Rev 29:737–775.
- 2555.Covaci A, Harrad S, Abdallah MA, Ali N, Law RJ, Herzke D, de Wit CA (2011) Novel brominated256flame retardants: A review of their analysis, environmental fate and behaviour. Environ Int25737:532–556.
- 258 6. Dodson RE, Perovich LJ, Covaci A, Van den Eede N, Ionas AC, Dirtu AC, Brody JG, Rudel RA
 259 (2012) After the PBDE Phase-Out: A Broad Suite of Flame Retardants in Repeat House Dust
 260 Samples from California. Environ Sci Technol 46(24): 13056-66.
- 7. Wu XM, Bennett DH, Moran RE, Sjödin A, Jones RS, Tancredi DJ, Tulve NS, Clifton MS, Colón
 M, Weathers W, Hertz-Picciotto I (2015) Polybrominated diphenyl ether serum concentrations
 in a Californian population of children , their parents , and older adults : an exposure
 assessment study. Environ Heal 14:1–11.
- 8. Man YB, Chow K., Man M, Lam JC, Lau F., Fung WC, Wong MH (2015) Profiles and removal
 efficiency of polybrominated diphenyl ethers by two different types of sewage treatment
 work in Hong Kong. Sci Total Environ 505:261–268.
- 268 9. Covaci A, Voorspoels S, de Boer J (2003) Determination of brominated flame retardants, with
 269 emphasis on polybrominated diphenyl ethers (PBDEs) in environmental and human samples
 270 a review. Environ Int 29:735–756.
- Portolés T, Sales C, Gómara B, Sancho JV, Beltrán J, Herrero L, González MJ, Hernández F
 (2015) Novel Analytical Approach for Brominated Flame Retardants Based on the Use of Gas
 Chromatography-Atmospheric Pressure Chemical Ionization-Tandem Mass Spectrometry with
 Emphasis in Highly Brominated Congeners. Anal Chem 87:9892–9899.
- 11. Megson D, Robson M, Jobst KJ, Helm PA, Reiner EJ (2016) Determination of Halogenated
 Flame Retardants Using Gas Chromatography with Atmospheric Pressure Chemical Ionization
 (APCI) and a High-Resolution Quadrupole Time-of-Flight Mass Spectrometer (HRqTOFMS).
 Anal Chem 88:11406–11411.
- 279 12. Feo ML, Barón E, Eljarrat E, Barceló D (2012) Dechlorane Plus and related compounds in aquatic and terrestrial biota : a review. Anal Bioanal Chem 2625–2637.
- Shen L, Reiner EJ, Macpherson KA, Kolic TM, Sverko E, Helm P., Bhavsar SP, Brindle ID, Marvin CH (2010) Identification and screening analysis of halogenated norbornene flame retardants
 in the Laurentian Great Lakes: Dechloranes 602, 603, and 604. Environ Sci Technol 44:760–
 766.
- Sverko E, Tomy GT, Marvin CH, Zaruk D, Reiner E, Helm PA, Hill B, McCarry BE (2007)
 Dechlorane Plus Levels in Sediment of the Lower Great Lakes. Environ Sci Technol 42:361–
 366.
- Sverko E, Tomy GT, Reiner EJ, Li YF, McCarry BE, Arnot JA, Law RJ, Hites RA (2011) Dechlorane
 plus and related compounds in the environment: A review. Environ Sci Technol 45:5088–
 5098.
- 16. Brasseur C, Pirard C, L'homme B, De Pauw E, Focant J-F (2016) Measurement of emerging
 dechloranes in human serum using modulated gas chromatography coupled to electron
 capture negative ionization time-of-flight mass spectrometry. Rapid Commun Mass Spectrom
 30:2545–2554.
- 295 17. De la Torre A, Concejero M, Sverko E, Shen L, Martínez M, Reiner E, Alaee M (2010) EFFECT OF

- SOURCE TEMPERATURE ON THE ECNI / MS SPECTRA OF DECHLORANE PLUS ISOMERS.
 Organohalogen Compd 72:1737–1740.
- Ren G, Yu Z, Ma S, Li H, Peng P, Sheng G, Fu J (2009) Determination of dechlorane plus in
 serum from electronics dismantling workers in South China. Environ Sci Technol 43:9453–
 9457.
- 301 19. Yan X, Zheng J, Chen KH, Yang J, Luo XJ, Yu LH, Chen SJ, Mai BX, Yang ZY (2012) Dechlorane
 302 Plus in serum from e-waste recycling workers: Influence of gender and potential isomer303 specific metabolism. Environ Int 49:31–37.
- 304 20. He S, Li M, Jin J, Wang Y, Bu Y, Xu M, Yang X, Liu A (2013) Concentrations and trends of
 305 halogenated flame retardants in the pooled serum of residents of Laizhou Bay, China. Environ
 306 Toxicol Chem 32:1242–1247.
- 30721.Zhou NS, Siddique S, Lavoie L, Takser L, Abdelouahab N, Zhu J (2014) Hexachloronorbornene-308based fl ame retardants in humans : Levels in maternal serum and milk. Environ Int 66:11–17.
- Cequier E, Marcé RM, Becher G, Thomsen C (2013) Determination of emerging halogenated
 flame retardants and polybrominated diphenyl ethers in serum by gas chromatography mass
 spectrometry. J Chromatogr A 1310:126–132.
- Cequier E, Maria R, Becher G, Thomsen C (2015) Comparing human exposure to emerging and
 legacy fl ame retardants from the indoor environment and diet with concentrations measured
 in serum. Environ Int 74:54–59.
- Fromme H, Cequier E, Kim JT, Hanssen L, Hilger B, Thomsen C, Chang YS, Völkel W (2015)
 Persistent and emerging pollutants in the blood of German adults: Occurrence of dechloranes,
 polychlorinated naphthalenes, and siloxanes. Environ Int 85:292–298.
- Poma G, Malarvannan G, Voorspoels S, Symons N, Malysheva S V., Van Loco J, Covaci A (2016)
 Determination of halogenated flame retardants in food: Optimization and validation of a
 method based on a two-step clean-up and gas chromatography-mass spectrometry. Food
 Control 65:168–176.
- 322 26. Dirtu AC, Dirinck E, Malarvannan G, Neels H, Gaal L Van, Jorens PG, Covaci A (2013) Dynamics
 323 of Organohalogenated Contaminants in Human Serum from Obese Individuals during One
 324 Year of Weight Loss Treatment. Environ Sci Technol 47:12441–12449.
- 325 27. Covaci A, Roosens L, Dirtu AC, Waegeneers N, Van Overmeire I, Neels H, Goeyens L (2009)
 326 Brominated flame retardants in Belgian home-produced eggs: Levels and contamination
 327 sources. Sci Total Environ 407:4387–4396.
- Barón E, Eljarrat E, Barceló D (2012) Analytical method for the determination of halogenated
 norbornene flame retardants in environmental and biota matrices by gas chromatography
 coupled to tandem mass spectrometry. J Chromatogr A 1248:154–160.
- 331

Compounds	рт	0		i	iLOD (fg μL ⁻¹)					
Compounds	RT	Q	q	2 μL	5 μL SV	10 µL SV				
BDE 28	8.05	79	81	20	10	5				
BDE 49	9.06	79	81	20	10	5				
BDE 47	9.26	79	81	10	5	2.5				
BDE 66	9.51	79	81	20	10	5				
BDE 100	10.36	79	81	20	10	5				
BDE 99	10.72	79	81	20	10	5				
BDEs 85	11.40	79	81	20	10	5				
BDE 154	11.69	79	81	20	10	5				
BDE 153	12.24	79	81	20	10	5				
BDE 138	12.95	79	81	20	10	5				
BDE 183	13.74	79	81	20	10	5				
BDE 209	21.04	487	489	200	100	50				
HBB	8.80	79	81	20	10	5				
TBB	10.65	357	359	200	100	50				
DEC602	10.70	612	35	250	120	60				
DEC603	13.30	638	35	100	50	25				
BTBPE	14.18	79	81	20	10	5				
ТВРН	14.60	384	515	100	50	25				
s-DP	14.90	654>35	654>37	1	0.5	0.25				
a-DP	15.22	654>35	654>37	0.5	0.25	0.125				
DBDPE	23.95	79	81	1000	500	250				

Table 1. Analytical performance of the method including MS quantitation parameters.

Table 2. Techniques and conditions previously used for the determination of DP isomers and their performance in terms of LOD and LOQ

N⁰	Technique	Column	Separation	m/z	LOD	LOQ	Ref
1 GC-ECNI-MS		DB-5 (15 m × 0.25 mm × 0.10 μm)	90 °C (1,5 min); 10 °C/min to 300 °C (3 min); 40 °C/min to 310 °C (5 min)	[M-H] ⁻ ;650, 652	n.a.	2 ng g ⁻¹ (DUST)	[6]
2	GC-APCI-HRqTOFMS	DB-5 HT (15 m × 0.25 mm × 0.10 μm)	110 °C; 40 °C/min to 200 °C; 10 °C/min to 280 °C; 30 °C/min to 330 °C (5 min)	[M] ⁻ ; 653.711	0.16 pg μL ⁻¹	n.a.	[11]
3	GC-EI-HRMS	DB-5 (15 m × 0.25 mm × 0.10 μm)	120 °C (1 min); 30 °C/min to 240 °C; 5 °C/min to 275 °C; 40 °C/min to 320 °C (3 min)	[M-C ₁₃ H ₁₂ Cl ₆] ⁺ ; 271.8102; 273.8072	0.5 pg g ⁻¹ (sediment), 15 pg g ⁻¹ (fish)	n.a.	[13]
4	GC-ECNI-MS	DB-5 (30 m × 0.25 mm × 0.25 μm) DB-35 (30 m × 0.25 mm × 0.25 μm) (confirmation)	80 °C (2 min); 10 °C/min to 285 °C (5 min)	[M-H] ⁻ ; 650.652	30 pg g ⁻¹ (sediment)	n.a.	[14]
5	CZC-GC/ECNI- TOFMS	Rtx- PCB (15 m × 0.25 mm × 0.25 μm) plus Rxi-17 (1 m × 0.18 mm × 0.18 μm)	140 °C (2 min); 30 °C/min to 280 °C; 5 °C/min to 300 °C (10 min)	[M-H] ⁻ ; 650, 652	3 pg (iLOD)	n.a.	[16]
6	GC-ECNI-MS	DB-XLB (30 m×0.25 mm×0.25 μm)	110 °C (1 min); 8 °C/min to 180 °C (1 min); 2 °C/min to 240 °C (5 min); 2 °C/min to 280 °C (15 min; 10 °C/min to 310 °C (5 min)	[M] ⁻ ; 653.8 and 651.8	n.a.	3.08 pg g ⁻¹ fat (serum) (s-DP), 1.29 pg g ⁻¹ fat (serum) (a-DP)	[19]
7	GC-ECNI-MS	DB-1MS (30 m × 0.25 mm × 0.25 μm)	120 °C (1 min), 10 °C/min to 300 °C (8 min); 10 °C/min to 310 °C (12 min)	[M] ⁻ ; 651.7 and 653.7	40 pg g ⁻¹ l.w. (s- DP), 120 pg g ⁻¹ l.w. (a-DP), (serum)	n.a.	[21]
8	GC-ECNI-MS	DB-5 (15 m × 0.25 mm × 0.10 μm)	50 °C, 25 °C/min to 300 °C (5 min)	[M] ⁻ ; 653.8 and 651.8	1.1 pg mL ⁻¹ (serum) (s-DP), 3.3 pg mL ⁻¹ (serum) (a-DP)	3.5 pg mL ⁻¹ (serum) (s-DP), 10 pg mL ⁻¹ (serum) (a-DP)	[22]

n.a. – not available

Compounds	LOD (pg g ⁻¹)	LOQ (pg g ⁻¹)	Biscuits	Smoked Salmon	Chicken Egg	Hyena Serum	Cheetah Serum	Lion serum	Human Cord Blood Pool	Blank	Chicken Egg	albumin	Chicken Egg	Sea Eagle Plasma	Sea Eagl Plasma
DEC602	1	2.5	<1	<1	<1	<1	66	12	<1	<1	5	<1	8	<1	14
DEC603	1	2.5	<1	<1	<1	8	19	8	<2.5	<1	<2.5	<1	<2.5	9	<1
syn-DP	0.03	0.10	<0.03	<0.03	<0.03	0.33	2.5	13	0.13	<0.03	9000	4	2680	1.95	<0.03
anti-DP	0.05	0.15	<0.05	<0.05	<0.05	<0.15	6	18	0.19	<0.05	29000	6	3450	22	<0.05
ΣDPs			<0.03	<0.03	<0.03	0.33	9	31	0.32	<0.003	38000	10	6128	24	<0.003
fAnti							0.71	0.58	0.59		0.76	0.4	0.56	0.92	
BDE 28	0.8	20	<0.8	<0.8	<0.8	<0.8	<20	<0.8	<0.8	<0.8	<0.8	<0.8	26	<0.8	<0.8
BDE 49	0.3	1	<0.3	35	<1	<0.3	<0.3	<1	<0.3	<1	220	<1	70	1	12
BDE 47	0.7	3	41	233	44	<3	11	4.5	<0.7	<0.7	540	6	455	3	16
BDE 66	1	10	<1	<1	<10	<1	<1	<1	<1	<1	<1	<10	46	<1	<1
BDE 100	0.4	4	9	51	6	<0.4	<0.4	4	<0.4	<0.4	440	7	270	4	4
BDE 99	0.5	2	18	7	5	8	8	2	<2	<0.5	1390	26	587	3	6
BDE 85	3	9	<9	9	<9	<3	<3	<3	<3	<3	20	<3	21	<3	<3
BDE 154	0.05	0.22	1	22	<0.22	3	56	9	0.22	<0.05	340	2	126	1.4	2.7
BDE 153	0.8	2	35	13	4	16	18	2	<0.8	<2	870	5	635	4	3
BDE 138	0.3	1	1	20	2	<0.3	13	42	<0.3	<0.3	50	<0.3	58	<0.3	<1
BDE 183	0.7	2	<0.7	<0.7	<0.7	<0.7	<2	<2	<0.7	<0.7	1900	4	1700	<0.7	3.5
BDE 209	6	20	92	84	92	58	610	<6	<6	<6	12000	55	6650	26	1138
ΣPBDEs			197	467	153	88	723	63	1.3	2	17800	105	10650	42	1185
DBDPE	13	30	<13	<13	<13	<13	<13	<13	<13	<13	<13	<13	<13	<13	<13
НВВ	5	15	<15	<15	<15	<5	<5	<5	<5	<5	250	<5	50	<5	<5
твв	10	40	<40	<10	<10	<10	<10	<10	50	<10	<10	124	<10	<10	<10
BTBPE	0.24	1	8	3	5	8	34	7.2	<1	<1	300	14	1990	4	4
тврн	5	15	72	32	53	<5	<5	<5	<5	<5	45	290	290	<5	159
ΣNBFRs			94	44	60	8	34	7.2	50		600	427	2325	4	163

Table 3. Concentrations of PBDEs and other HFRs (pg g^{-1}) in the analyzed samples.

*< xx: below the respective LOQ or LOD

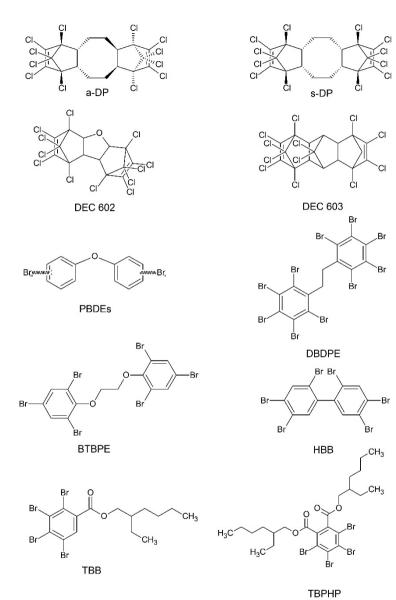
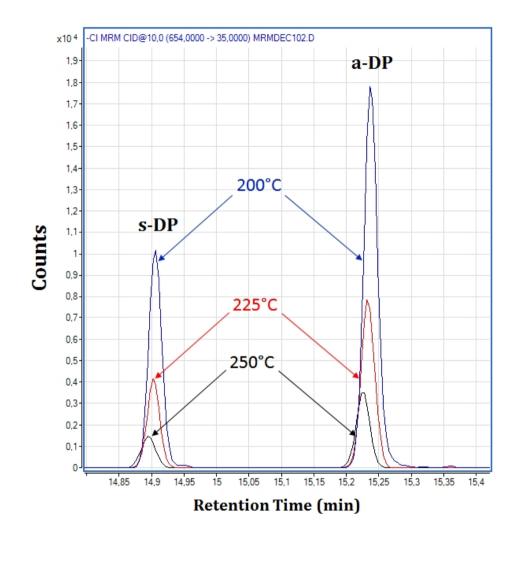


FIG 1.



343 FIG 2

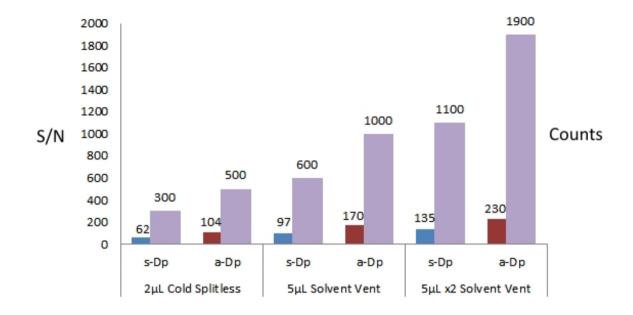


FIG 3.

