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1 **Assessment of PM10 and heavy metals concentration in a Ceramic**
2 **Cluster (NE Spain) and the influence on nearby soils.**

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8
9 **Abstract**

10
11 Environmental pollution control is one of the most important goals in pollution risk
12 assessment today. The aim of this study is conducting a retrospective view of the
13 evolution of matter particulate (PM10) and the heavy metals (Cd, Ni and Pb) at different
14 localities (Alcora, Castellón and Onda) in the Spanish cluster ceramic in a period
15 between January 2007 and December 2011. The study area is in the province of
16 Castellón. This province is a strategic area in the framework of European Union
17 Pollution control. Approximately 80% of European ceramic tiles and ceramic frits
18 manufacturers are concentrated in two areas, forming the so-called “Ceramics Clusters”;
19 one is in Modena (Italy) and the other in Castellón (Spain). In these kind of areas, there
20 are a lot of pollutants from this industry that represent an important contribution to soil
21 contamination so it is necessary to control their air quality. In these areas atmospheric
22 particles are deposited in the ground through both dry and wet deposition. Soil is a
23 major sink for heavy metals released into the environment. For this purpose the levels of
24 PM10 in ambient air and the corresponding annual and seasonal trend were calculated.
25 The results of the study show that the PM10 and heavy metals concentrations are below
26 the limit values recommended by European Union Legislation for the protection of

27 human health and ecosystems in the study period. There is an important reduction of
28 them from 2009 in all control stations due to economic crisis and subsequent decrease
29 of industrial activity. The atmospheric seasonal tendency of pollutants concentrations is
30 marked by the rate of industrial activity and additionally by the temperature.
31 Complementary, a comparative study of heavy metals levels in soils was performed in
32 this area. Soils with low pollution by Ni and Pb were detected, while different pollution
33 by Cd was found depending on the sampling site. Although there is an evident reduction
34 of PM10 and heavy metals levels, the results show that these pollutants have been
35 accumulated in the soil close to emission sources.

36

37

38

39 Keywords: PM10, Heavy metals, soil pollution, ceramic cluster

40

41 **1. Introduction**

42

43 Environmental pollution control is nowadays one of the most important goals in
44 pollution risk assessment. Metal contamination of the environment raises concern for
45 the possible impact on human health, and the existence of heavy metals in soils, of both
46 natural -inherited from the origin material (Galán et al., 2008), lithogenic or pedogenic;
47 (Tume et al., 2011)- or anthropogenic origins -due to human activities as agricultural
48 projects, water waste discharges or atmospheric emissions (Gray et al., 2003; Bech et
49 al., 2008; Hovmand et al., 2008)-, is well-recognized as a potentially important source
50 of human exposure (Muller and Anke, 1994; Davis et al., 2009; Chen et al., 2005).

51

52 PM10 particles and heavy metals are emitted into the atmosphere as aerosols, mainly as
53 a result of human activities, and are deposited on the ground (Borgna et al., 2009). So
54 the deposition of atmospheric particles is an important source of soil contamination in
55 areas exposed to significant pollution levels. Sakagami et al. (1982) reported that there
56 was a close relationship between heavy metal concentration in soils and those in the
57 airborne particles that fall. Other authors have found that atmospheric inputs of heavy
58 metals to agricultural systems can be a significant contribution to metal loading in soils.
59 For example, Alloway. (1999) found a mean atmospheric deposition rate for Cd of
60 around $1.9 \text{ g ha}^{-1}\text{y}^{-1}$, which Nicholson et al. (1999) calculated to be 50% of the total
61 annual input of Cd to agricultural land in the UK. Nicholson et al. (1999) also calculated
62 that atmospheric deposition of other metals such as Ni to be between 32-45% of the
63 total annual inputs of these metals to soils. Then, in comparison to the other forms of
64 metals inputs, potentially, atmospheric inputs have much more immediate impact on
65 agricultural systems (Gray et al., 2003).

66

67 Once contaminated, soils typically remain in this condition for protracted periods of
68 time because of sorption of metals on to particles and limited mobility (Peris et al., 2008;
69 Ferri et al., 2012). This fact produces a potential human cumulative exposure. The
70 accumulation of heavy metals in agricultural soils, including home vegetables gardens
71 may be of particular concern since consumption of vegetable grown in metal
72 contaminated soils may pose health risks for the population residing in these areas (Cui
73 et al., 2005; Intawongse and Dean, 2006). Elevated concentrations of trace elements in
74 the soil–water–plant ecological system are of great concern because of possible
75 influences on the food chain. (Tume et al., 2008). While essential trace metals, such as
76 Ni, are necessary for plant growth and/or human nutrition at low levels, they may also
77 be toxic to both animals and humans at high exposure. Other trace elements, for
78 example Cd and Pb, may also inadvertently enter the food chain and pose health risks to
79 human and animals (Laughlin et al., 1999, Micó et al., 2006). Prolonged metal exposure
80 spanning neurodevelopmental periods may also increase the risk of neurodegenerative
81 conditions for elderly people (Aelion et al., 2008). Moreover, heavy metals in soils can
82 generate airborne particles which may affect the air environmental quality (Gray et al.,
83 2003).

84

85 The aim of this study is conducting a retrospective view of the evolution of particulate
86 matter (PM₁₀) and selected heavy metals (Cd, Ni and Pb) at different localities
87 (Castellón, Alcora and Onda, fig.1) in the Spanish cluster ceramic from January 2007 to
88 December 2011. Complementary, a comparative study of heavy metals levels in soils
89 was performed in this area. The data were extracted from the papers of Roca-Pérez et al.,
90 (2010) and Jordán et al., (2009).

91

92 **2. Description of the study area**

93

94 The study area is located in the East of Spain, in the province of Castellón. This
95 province is a strategic area in the frame work of the European Union (EU) pollution
96 control. Approximately 80% of European ceramic tiles and ceramic frit manufacturers
97 are concentrated only in two areas, forming the so-called “ceramics clusters”; one is in
98 Modena (Italy) and the other in Castellón.

99

100 The type of climate in the study area is Mediterranean characterized by wet and mild
101 winters, dry and warm summers, and an average temperature variation of 13.5°C.
102 Rainfall is abundant in spring and autumn, coinciding with the dominance of western
103 winds while summer is drier, dominated by the Azores anticyclone (Vicente et al.,
104 2011). Yearly rainfall generally does not exceed 400 mm.

105

106 This area has a complex Mediterranean atmospheric environment, with low rainfall, soil
107 with poor vegetation coverage and frequent high particulate air-mass intrusions from the
108 Sahara (Rodríguez *et al.*, 2002). A system of local sea breezes is also present in the
109 study area due to geographical characteristics and the proximity to the sea. These
110 periodic land-sea winds, which have been extensively studied by several authors
111 (Martín et al., 1991; Boix *et al.*, 1995; Millán *et al.*, 2001; Sanfeliu et al., 2002), govern
112 the microclimate in this area, resulting in an overall effect of smoothing the
113 temperatures (Pogosyan, 1965). Due to this system of breezes, the concentration of
114 pollutants may be affected by emission sources located outside the study area on a daily
115 basis (Fig. 1).

116

117 The origin of air PM₁₀ in this area is both natural and anthropogenic. The former is due
118 to the resuspension of mineral materials from the surrounding mountains with poor
119 vegetation coverage and the long-range transport of materials from North Africa
120 (Rodríguez *et al.* 2001, Pérez *et al.* 2006; Esteve *et al.* 2012). These dust intrusions
121 from North Africa influence ambient PM₁₀ levels in the study area at around 2 µ/m³ on
122 an annual basis (Minguillón *et al.*, 2009).

123

124 Anthropogenic pollution sources originate from automobile traffic (mobile sources) and
125 industrial activity (fixed sources). The main industrial activity in the study area is based
126 on producing ceramic tile (Vicente *et al.* 2007). This industrial sector has two types of
127 factories, one for the manufacture of tiles and the other to supply raw materials. The raw
128 materials of the tile body consist mainly of clay from sources such as opencast quarries
129 within the ceramic cluster area (Jordán *et al.*, 2009; Sanfeliu *et al.*, 2009). The raw
130 materials for decoration involve manufacture frits, enamels, and colour (Jordán *et al.*,
131 2006). In the manufacture of ceramic tile, channelled and diffuse emissions from the
132 production processes and the storage, handling and transport of raw materials all
133 increase the concentration of particles in the air (Sanfeliu *et al.*, 2002). However,
134 particle emissions from the manufacture of pigments, frits and enamels probably have a
135 greater impact on the levels of heavy metals than on particle mass. (Minguillón *et al.*,
136 2007). An additional important factor is that a power station, a refinery and several
137 chemical industries are located at east of the study area (Boix *et al.*, 2001) (Fig. 1).
138 These industries together contribute to environmental pollution in the area. Finally,
139 relevant sources of secondary PM in the area include precursor emissions of the volatile
140 organic compounds (VOC's), NO_x and SO₂ from high temperature ceramic processes,

141 power generation, petrochemical processes and biomass combustion (Minguillón *et al.*,
142 2007).

143

144 In the case of chemical pollutants in air PM₁₀, nickel is found as a trace element in
145 petrol, and therefore its release into the atmosphere is related mainly to the combustion
146 of fossil fuels (coal and fuel oil) in electricity and heat production and also in traffic
147 exhausts (Pacyna *et al.* 1984; Ghio and Samet 1999). Nickel oxides are also widely used
148 as components of pigments used in the ceramics industry. Concentration levels of
149 cadmium in ambient air are associated with industrial processes in the manufacturing of
150 frits and enamel. Emissions of cadmium are also produced in the processes of a nearby
151 power station (Boix *et al.* 2001). The most important emission of lead are traffic
152 exhausts. Petrol additives contain lead (Parekh *et al.* 2002), which after combustion is
153 released into the atmosphere as organic lead (lead bromide and lead chlorinebromide)
154 (Pacyna 1998). With the introduction of new international laws, the use of lead in petrol
155 has been banned, and this contribution is now minimal, its use reduced to obsolete
156 means of transportation. In the ceramics industry, lead oxides are also used extensively
157 as a component of pigments. Relationships between the emissions from this sector to air
158 ambient levels of lead in close urban areas have been identified (Sanfeliu *et al.* 2002;
159 Gómez *et al.* 2005).

160

161 The study is focused on 154 km² that is the surface between the three cities studied.
162 Alcora is situated at 279 m above sea level, Onda at 194 m and Castellón at 30 m. There
163 is a high altitude difference between these cities in a small area. The distance between
164 Castellón, Alcora and Onda is 17 km.

165

166 The study area is based on Quaternary materials generated by torrents from inland
167 massifs (Maestrazgo Range and tributaries from the Sierra de Espadán range), which
168 have formed lagoons closer to the sea. This process is reported to have started at the end
169 of the Tertiary period, as the Pliocene became the Pleistocene era, favoring a climate
170 suitable for torrential erosion and subsequent transport and deposition. The main source
171 of sediments has decreased in the last few years because of regulations of the course of
172 the river (Vicente et al., 2007)

173

174 Alcora is in the Miocene surrounded by Cretaceous and Quaternary. Its materials are
175 conglomerates, limestone, marl, clay and lignite. The materials around this city are
176 limestone, marl, sand, clay and gravel. Onda is located in the Quaternary enclave,
177 surrounding a Jurassic area. This town is surrounded by Quaternary and Triassic. The
178 main materials of this city are conglomerates, limestones and dolomites, around marl,
179 clay and limestone in the Triassic and red clays in the Quaternary. Castellón is located
180 in the Quaternary and the main materials are clays, silts and sands. All deposits are
181 Quaternary sedimentary.

182

183 **3. Materials and Methods**

184

185 3.1. Atmospheric particles

186 *Sampling collection*

187 A PM10 medium volume sampler model IND-LVS3 manufactured by KleinfILTERGERÄT
188 was used. This device is considered as a reference according to European regulations
189 (European Council Directive 2008/50/EC; UNE-EN 12341:1999), for the sampling of
190 PM10 particles. The technology used in the equipment consists of blowing air through

191 an inlet with a vacuum pump. The particulate matter was blown in through the opening
192 circumference between the frame and the round cover mounted on top. Within the
193 sampler inlet the airflow was accelerated by eight impactor nozzles and then directed
194 toward the impacting surface. Particles were trapped on a permeable support consisting
195 of a 47 mm diameter filter. The device contains a temperature sensor with a radiation
196 protector that eliminates deviations in the reading caused by solar radiation in addition
197 to a pressure sensor. The sampling flow volume was 2,3 m³/h during 24 h periods. A
198 total of 4.100 PM10 samples (1.496 Castellón, 1.253 Alcora and 1.351 Onda) were
199 collected in filters from 2007 to 2010. The filters used were quartz fiber filters were
200 used, according to UNE-EN 12341:1999. They were made from SiO₂ pure base, totally
201 free of additives. These filters allow an efficiency of separation greater than 99.5%.

202

203 The samplers of the PM10 were positioned between 1,5 and 4m about the ground
204 according to the Directive 2008/50/EC (Annex III, Microscale siting of sampling
205 points).

206

207 *Gravimetric analysis*

208 Particle concentrations levels were determined gravimetrically. This method consists of
209 weighting the filters twice: firstly empty and then with sample. The filters must be kept
210 for at least 48h in a special chamber. The conditions inside the chamber are 50%
211 relative humidity and a temperature of 20°C, according to normative UNE-EN
212 12341:1999. Filters were weighted on an analytical balance with a precision of 0.1 mg.
213 The PM concentration levels were determined based on the sample quantities obtained
214 and the volume of air pumped.

215

216 *Chemical analysis*

217 The levels of Cd, Ni and Pb in the PM10 samples were determined by inductively
218 coupled plasma mass spectrometry (ICP-MS). The equipment used was an Agilent
219 model 7500CX that included a quadrupole, a collision cell and an integrated
220 autosampler. The equipment was installed in a chamber with a clean air filter unit and
221 an independent air conditioning system. This instrumental technique allows the Cd, Ni
222 and Pb levels to be rapidly identified after dissolution of the sample. Dissolution was
223 achieved by acid digestion in hermetic Teflon recipients. This methodology has been
224 used by many authors (Kubilay and Saydam, 1995; Querol *et al.*, 2000).

225

226 In order to detect any possible traces of contamination-causing As, Cd, Ni or Pb
227 contained in the reagents and quartz filter fibres, digestions with only reagents (blank
228 reagents) and filters without a sample (blank filters) were performed. The SRM 1648
229 “urban particulate matter” pattern was used to validate the results. This pattern consists
230 of particulate matter of anthropogenic origin collected in an industrialised urban
231 atmosphere and was an adequate standard of reference for this study.

232

233 3.2. Soil

234 *Sampling collection*

235 The sampled soils were selected between the different samples from the papers of Roca-
236 Pérez *et al.*, (2010) and Jordán *et al.*, (2009). The soils were chosen for their proximity
237 of the industrial atmospheric emission sources and their physicochemical property
238 diversities, especially pH (between 6.5 and 8) and carbonate content, the soil of Alcora
239 is noncalcareous while Castellon and Onda are calcareous. In Alcora, almond trees are
240 the main crop, followed by olive trees, vineyards, some carob trees, and a few fig trees.

241 In Castellón and Onda, the crop of orange trees is the main. These soils have a low
242 capacity use and a erosion risk. In addition in Onda, there are other terraced fields
243 destined for the cultivation of carob, olive, hazel and almond trees. All studied soils are
244 anthrosols, according to the FAO World Reference Base for Soil Resources. These soils
245 are a type formed or heavily modified due to long-term human activity, such as from
246 irrigation, addition of organic waste or wet-field cultivation used to create paddy fields.

247

248 According to the soil classification of FAO/UNESCO 1998, the soils chosen of the
249 papers of Roca-Pérez et al., (2010) and Jordán et al., (2009) are classified as:

250

251 - *Fluvisols*: Alcora and Castellón. Typical soils found of valley bottoms and
252 floodplains, widely represented in the province of Castellón. These occupy the
253 coastal plains and the terraces of the river courses from the inland areas to the
254 coast. However, inland areas may develop on slopes fitted out by terraces. These
255 soils are partly immature and they are commonly well drained. Because of their
256 development from recent alluvial deposits, these soils have a significant
257 thickness and a variable texture (Antolín Tomás, 1998).

258 - *Calcisol*: Onda. These types of soils are characterized by horizons with
259 secondary enrichment of calcium carbonate, favored by semi-arid conditions.
260 They are present either in the coastal plain or in the mountainous interior. In
261 these soils predominate washing and carbonate accumulation is associated with
262 geomorphic processes of glaciais formation, alluvial fan, etc. (Antolín Tomás,
263 1998).

264

265 These authors collected two kilograms of topsoil samples (0-20 cm) from each sampling
266 site, air-dried, crushed, 2 mm sieved, mixed and stored at air-dried conditions for further
267 analysis.

268

269 *Chemical analysis*

270 The methodology used in soil chemical analysis was described by Roca-Pérez *et al.*
271 (2010) and Jordán *et al.* (2009) in their studies. Total contents of metals were
272 determined by ICP-MS after microwave digestion of 0.5g of representative soil sample
273 using 9 ml nitric acid (68% w/v), 1 ml hydrogen peroxide (30% w/v), 3 ml hydrofluoric
274 acid (48% w/v), 2 ml hydrochloric acid (37% w/v) and 5 ml deionised water in the first
275 steps (20 min at 200°C) and 30ml boric acid (4%) in the second (5min at 170°C) (EPA,
276 1996).

277

278 **4. Results and discussion**

279

280 4.1. Atmospheric concentrations of pollutants.

281 Table 1 shows the assessment of PM₁₀, Cd, Ni and Pb according the limit values
282 established by European Legislation (European Council Directive, 2008/50/EC). The
283 concentrations of the pollutants are below the limit values recommended by European
284 Union Legislation for the protection of human health and ecosystems in the study period,
285 with the exception of daily exceedances of PM₁₀ in 2007 in the station of Castellón.

286

287 There is an important reduction of the pollutants from 2009 in all control station due to
288 the economic crisis and subsequent decrease of industrial activity. Additionally, due to
289 the new European Directive (2010/75/EU) on industrial emissions, the best available

290 techniques (BAT) were implemented in the industry in order to prevent or reduce
291 harmful impacts.

292

293 In Modena ceramic cluster this reduction has been also observed despite the Italian
294 ceramic industry was the first in Europe to develop and adopt techniques for the
295 reduction of the environment impacts in the seventies (Minguillón *et al.*, 2013). So the
296 economic crisis has been very important in the two industrial sites.

297

298 4.1.1 PM10

299 In general, PM10 levels increase during the months with high temperatures due to a
300 decrease in precipitation (Fig. 2) with the exception of 2007 and 2008 in Castellón and
301 Alcora. This causes a reduction in the cleansing effect on the atmosphere (Bergametti *et*
302 *al.* 1989) and consequently a greater contaminant concentration in the ambient air. The
303 high temperatures during these months lead to increased dryness of the terrain, which
304 favours the resuspension of clay-loam substrate in the area (Gómez *et al.*, 2005). At the
305 same time, the mixing layer, or lower part of the troposphere where the pollutants are
306 free to move through the turbulence generated in the lower layers of the atmosphere,
307 increases its thickness and facilitates the mixing of air masses from the North of Africa
308 in the low layers (Kubilay and Saydam, 1995). Intrusion episodes of long-distance
309 material occur, leading to an increase in the concentration of PM10. During winter
310 months, temperature inversions are generated. This phenomenon occurs on clear nights,
311 with weak or no wind, when the soil losses the heat acquired by radiation and low-lying
312 air layers are cooled faster than the upper layers of air (Wallace *et al.*, 2010). When
313 pollutants are emitted under temperature inversion conditions, they accumulate in the
314 layers of the troposphere close to the ground. This phenomenon causes transport

315 through these layers to occur too slowly, producing an increased concentration of
316 pollutants (Monn et al., 1995). This accumulation of pollutants is also found in Milan
317 due to persistent thermal inversions (Marcazzan *et al.*, 2001). During the autumn season,
318 the lowest values of the study were detected (Fig. 2). This was due to atmospheric
319 instability, the tendency of the atmosphere to resist or enhance vertical motion or,
320 alternatively, to suppress or augment existing turbulence (Zoras *et al.*, 2006). According
321 to some studies, as global weather conditions change, the input frequency of air masses
322 from North Africa is reduced (the mixed layer decreases), rainfall increases, and there is
323 a greater cleansing effect in the atmosphere (Querol *et al.*, 2002, 2004).

324

325 The seasonal evolution of the pollutants concentrations could supply valuable
326 information about the potential origin of them. These seasonal variations are dominated
327 by changes in meteorological conditions (Chang and Lee, 2008) and by human activities
328 (Guangjian et al., 2009).

329

330 4.1.2. Heavy Metals

331 The constant tendency of de cadmium concentration during the year (Fig. 3) indicates
332 that the contribution of this pollutant into the atmosphere is continuous from all sources,
333 both natural and anthropogenic. The small variation is due to one of the principal
334 origins of cadmium in the study area, the use of this element in the form of oxide in the
335 formulation of enamels and pigments in the ceramic industry (Matthes 1985). This
336 sector cuts its production levels in the summer months and so cadmium levels were
337 accordingly lower. In addition, the power plant located within the study area also emits
338 cadmium (Boix et al. 2001), which is higher in the months of higher energy demand
339 (cold months), wherein a light cadmium concentration increase was observed.

340

341 The light seasonal tendency for nickel (Fig. 4) of high concentrations during hot months
342 and low ones during the coldest months leads to nickel concentration levels being
343 governed by two factors: anthropogenic activities within the study area and the ambient
344 temperature. High-temperature anthropogenic activities at the power and petrochemical
345 plants, etc., release nickel into the atmosphere mainly as fly ash and not in a gaseous
346 state (Boix et al. 2001). Due to their characteristics, these fine particles behave in the
347 atmosphere like a gaseous pollutant (Wark and Warner, 2000), and so their dispersion is
348 influenced by external factors like the temperature (Vicente et al., 2011). Higher
349 temperatures result in greater atmospheric turbulence (Wark and Warner, 2000),
350 bringing greater dispersion away from the emitting sources located to the east of the
351 study area. The power and petrochemical plants are located there, thus favoring the
352 enrichment of nickel in the particle concentration levels in farther-away zones during
353 the hottest months.

354

355 The general behavior of lead concentration (Fig. 5) is to reach high levels in colder
356 seasons and low levels during warmer seasons. In the study area, the origin is linked to
357 the use of lead in the formulation of raw materials in the ceramic sector (Matthes 1985).
358 These oxides are volatilized in high-temperature industrial processes and they condense
359 upon contact with the atmosphere. An aerosol–vapor equilibrium exists for lead
360 dependent upon the air temperature. Low temperatures favor the aerosol phase, while
361 high temperatures favor the vapor phase release, explaining that the dispersion is greater
362 at high temperatures, and therefore lead levels are less in areas near the emitting
363 sources. Dispersion is less at low temperatures and so greater concentrations of this
364 contaminant abound near the emitting sources (Vicente et al., 2011). Another factor

365 needing consideration is that, in summer, this sector reduces production during vacation
366 periods, resulting in less emissions and lower lead concentration levels in the ambient
367 air.

368

369 4.2 Soil pollution.

370

371 Heavy metal concentrations in some different environments show high variability
372 depending on the type of environment (industrial, urban or natural). Some authors have
373 conducted studies of soils in different environments (Galán et al., 2002; Moral et al.,
374 2005; Tong-Bing et al., 2005; Yay et al., 2008; Maas et al., 2010) and concluded that
375 the soils most polluted by heavy metals occur in areas affected by mining and cement
376 factories (Soriano et al., 2012).

377

378 Soils affected by heavy metal emissions and PM10 in the ceramic cluster are carbonate
379 reliefs, detrital or gypsiferous materials subsequently colluvial-alluvial deposits and
380 marsh. The presence of carbonate indicates high soil pH, which tends to precipitate the
381 heavy metals. The Cd, and other metals have a strong tendency to be adsorbed by
382 carbonates. The Cd mobility is medium in basic and neutral soils, low in Pb and very
383 low in Ni (García et al., 2009).

384

385 Table 2 shows classification of soil contamination types based on their heavy metals
386 concentrations, with four categories identified.

387

388 Table 3 shows heavy metals concentration in soil in the study area. According to the
389 polluted soil classification (table 2), in Onda and Alcora, the soils show no pollution by

390 nickel but in Castellón shows light pollution. In the case of cadmium, Onda has light
391 pollution, Alcora medium and in Castellón very extreme pollution was detected. The
392 soils from Alcora and Castellón are unpolluted by lead and Onda shows medium
393 pollution.

394

395 According to studies by D'Elia et al., (1999) the soils of the Bologna area have standard
396 values of lead as in Castellón and Alcora, low pollution of cadmium as in Onda, while
397 medium pollution of nickel were detected.

398

399 **5. Conclusions**

400

401 In this study a retrospective view of evolution of PM10 and the heavy metals (Cd, Ni
402 and Pb) was conducted at three cities in the Spanish ceramic cluster, from 2007 to 2011.
403 The results show that the concentrations of studied pollutants are below the limit values
404 of European laws for the protections of human health and ecosystems, with the
405 exception of daily exceedances of PM10 in 2007 in the station of Castellón. An
406 important reduction of the pollutants from 2009 in all control station are detected due to
407 the economic crisis and subsequent decrease of industrial activity

408

409 The atmospheric seasonal tendency of the pollutant concentrations in the study area is
410 marked by the rate of industrial activity and, additionally, by the ambient temperature.

411

412 Additionally, a comparative study of heavy metals in soils was performed in this area
413 because although the particle deposition process cleans the atmosphere, its ultimate
414 result is the transfer of toxic atmospheric pollutants into the soil. The study of metals

415 contents in soil is currently necessary to obtain reference values and assess their
416 contamination, because the pollution from atmospheric particulate metals in the soil
417 increases their toxicity

418

419 Soils with low pollution by nickel and lead were detected in the study area, while
420 different pollution by cadmium was found depending on the sampling site, Castellón
421 shows extreme pollution, Alcora medium and Onda low.

422

423 Even if there is an evident reduction of PM10 and heavy metals atmospheric levels in
424 the study area, these pollutants have been accumulated in the soil close to emission
425 sources for years. The persistence of these pollutants in soil is much longer than in other
426 compartments of the biosphere, especially heavy metals which can be considered
427 virtually permanent. So, it is very important to know and reduce the emissions in order
428 to minimize harmful effects on human health, with particular attention to sensitive
429 populations, and diminish damage to the environment as a whole. In this paper, the
430 importance of soil composition has been identified to develop environmental control
431 strategies

432

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434

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437

438

439

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